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Title:

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Date:

2021-03-01

Citation:

Burton, J. E., Cawson, J. G., Filkov, A. I. & Penman, T. D. (2021). Leaf traits predict global patterns in the structure and flammability of forest litter beds. *Journal of Ecology*, 109 (3), pp.1344-1355. <https://doi.org/10.1111/1365-2745.13561>.

Persistent Link:

<https://hdl.handle.net/11343/276723>

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Article type : Research Article

Handling Editor: Hans Cornelissen

Title

Leaf traits predict global patterns in the structure and flammability of forest litter beds

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This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as [doi: 10.1111/1365-2745.13561](https://doi.org/10.1111/1365-2745.13561)

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Abstract

1. Fallen plant material such as leaves, needles and branches form litter beds which strongly influence fire ignition and spread. Traits of the dominant species influence litter flammability directly by determining how individual leaves burn and indirectly through the structure of the litter bed. However, we are yet to determine the relative importance of these different drivers across a range of plant species from different biomes.
2. We undertook a meta-analysis, combining leaf trait, litter structure and flammability data for 106 species from North America, South America, Europe, Asia and Australia. The dataset encompassed broad-leaved and coniferous species from 7 different experimental studies. Relationships between leaf traits, litter structure and key flammability metrics – sustainability, combustibility and consumability – were analysed using bivariate and piecewise structural equation modelling (SEM).
3. Traits which characterise the three-dimensional nature of the leaf and how much space a leaf occupies showed much stronger associations to litter structure and flammability than other morphological traits. Leaf curl, surface area to volume ratio (SAV) and specific leaf area (SLA) predominately influence litter flammability indirectly via litter structure with SLA being the only leaf trait which had a negative direct effect on flame duration. Packing ratio and bulk density were influenced by different combinations of leaf traits and, in turn, they aligned with different flammability metrics. Bulk density predicted flame spread rate and flame duration whereas packing ratio predicted consumption.
4. *Synthesis.* We identified key leaf and litter traits which influence different components of litter bed flammability. Importantly, we show that the effects of these leaf and litter traits are consistent across a wide-range of taxa and biomes. Our study represents a significant step towards developing trait-based models for predicting

56 surface wildfire behaviour. Such models will more flexibly accommodate future shifts
57 in the composition of plant species and leaf traits triggered by altered fire regimes and
58 climate change.

59 *Keywords:* leaf traits, litter properties, fire behaviour, flammability, meta-analysis, surface
60 fuel.

61 **1. Introduction**

62 Wildfires are a major disturbance that shape the distribution and structure of plant
63 communities across the globe (Bond and Keeley, 2005). Plants are not only affected by
64 wildfire but, as fuel, they influence how wildfires behave (Schwilk and Caprio, 2011, Pausas
65 et al., 2017). Plant species differ in their flammability (ability to burn) as a function of their
66 physical and chemical traits (Pausas et al., 2017, White and Zipperer, 2010). Whilst the
67 number of studies examining relationships between plant traits and flammability is growing,
68 the generality of these relationships has not been evaluated.

69 Identifying the underlying traits responsible for interspecific variation in flammability is
70 important. Models that incorporate plant traits can be used to predict fire behaviour (Zylstra
71 et al., 2016, Morvan and Dupuy, 2004) and evaluate vegetation management strategies
72 (McColl-Gausden and Penman, 2019). Studies that rank species flammability can be used to
73 identify low flammable species for planting in fire-prone areas (Dimitrakopoulos and Panov,
74 2001, Krix et al., 2019, Murray et al., 2018). Additionally, knowledge linking plant traits to
75 flammability can help explain the evolution and function of plant species in fire-prone
76 environments (Pausas and Moreira, 2012, Pausas et al., 2017) and provides insight into how
77 changes in the distribution and abundance of plant species may affect future wildfire
78 behaviour (Babl et al., 2020, Stevens et al., 2020).

79 Recent studies have quantified interspecific variation in flammability at the leaf-level and
80 shoot-level for a range of plant species (Alam et al., 2020, Cui et al., 2020). However, an
81 equivalent examination of patterns at the litter bed-scale has not yet been made. Identifying
82 what drives flammability at different scales is critical as the results from one scale may not be
83 transferable to another scale (Ganteaume, 2018, Grootemaat et al., 2017a). This is
84 particularly important for litter beds which contain senesced, fallen leaves rather than green
85 leaves. In litter beds, the effect of plant traits on flammability is mediated by trait afterlife
86 effects (Cornelissen et al., 2017). Understanding trait afterlife effects on litter flammability is
87 especially important as most wildfires ignite and initially spread in the litter bed (Curt et al.,

88 2011, Burrows, 2001). Moreover, a fire in the litter bed promotes and sustains burning of
89 upper fuel layers, allowing fire to spread vertically (Plucinski et al., 2010).

90 Litter flammability is multidimensional and encompasses four components - ignitability,
91 sustainability, combustibility and consumability – which associate with different aspects of
92 fire behaviour (Martin et al., 1994, Anderson, 1970). Each of these components are not
93 necessarily positively related and may be driven by different traits or factors (Varner et al.,
94 2015). Ignitability refers to the ability of a litter bed to ignite and spread a certain distance
95 and is commonly measured as ignition success or time to ignition once exposed to a heat
96 source (Anderson, 1970). Sustainability describes how long litter continues to burn once
97 ignited and is commonly measured as the duration of flaming combustion (Martin et al.,
98 1994). Combustibility represents how rapidly or intensely litter burns, generally measured as
99 the flame spread rate (Martin et al., 1994). Finally, consumability captures how much litter is
100 consumed (Martin et al., 1994).

101 Studies linking leaf and litter bed traits to flammability have been conducted in different fire-
102 prone ecosystems across the world using mostly experimental laboratory approaches (Scarff
103 and Westoby, 2006, Cornwell et al., 2015a, Grootemaat et al., 2017a, de Magalhães and
104 Schwilk, 2012a, Parsons et al., 2015, Zhao et al., 2016, Engber and Varner, 2012, Fonda,
105 2001) but also some field-based approaches (Schwilk and Caprio, 2011). These studies
106 focused on the flammability of leaf litter beds, owing to the dominance of leaves in the litter
107 bed and the propensity of leaves to ignite due to their fineness and high surface area (Zhao et
108 al., 2019). Each study focused on local species or species from a single forest biome. To date,
109 no study has quantitatively evaluated whether the trends observed in these individual studies
110 are globally consistent across a range of plant species, encompassing both conifers and broad-
111 leaves.

112 A range of leaf morphological traits which are important to flammability have been identified
113 in these studies (Table 1). We focus only on morphological traits as they have been found to
114 be a stronger driver of litter bed flammability in experimental manipulations (Scarff and
115 Westoby, 2006, Cornwell et al., 2015a), although we acknowledge leaf chemistry (e.g.
116 terpene content) may also be important (Ormeño et al., 2009, Dewhirst et al., 2020, Romero
117 et al., 2019). Leaf size and shape influence flammability indirectly via litter structure and thus
118 litter bed aeration (Scarff and Westoby, 2006). Additionally, leaf size and shape can influence
119 litter bed flammability directly by determining how individual leaves burn and so how

120 quickly fire can spread from leaf-to-leaf in the litter bed (Martin et al., 1994). Thinner leaves
121 with higher surface area to volume ratio (SAV) and specific leaf area (SLA) are quicker to
122 ignite, which can lead to faster fire spread and shorter burning times (Grootemaat et al., 2015,
123 Murray et al., 2013, Rothermel, 1972).

124 Litter structure is a well-established driver of litter fire behaviour, influencing fire spread rate
125 and combustion duration (Rothermel, 1972). Various methods are used to report litter
126 structure such as bulk density (mass of litter per litter bed volume), packing ratio (volume of
127 litter per litter bed volume) or porosity (the fraction of void space in the litter bed). Multiple
128 traits related to the size and shape of leaves have been found to influence both the packing
129 ratio and bulk density of litter beds by different studies (Table 1). For example, larger (as
130 quantified by size, SLA, SAV) and curlier leaves create a more aerated structure (lower
131 packing ratio) which at low moisture contents burns more rapidly and leads to greater
132 consumption (Engber and Varner, 2012, Grootemaat et al., 2017a). A variety of leaf traits
133 have been found to be associated with bulk density, including length, leaf curl, specific leaf
134 area, thickness and leaf area (Table 1).

135 Exploring the patterns between leaf traits, litter structure and litter flammability provides
136 fundamental information for understanding how plant species influence wildfire behaviour.
137 In this study, we combined the results of several key experimental studies on litter
138 flammability and asked: how do plant species' leaf traits determine litter bed properties and
139 thereby flammability across a wide range of taxa from different biomes?

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150 **Table 1. Summary of morphological leaf traits and their influence on litter bed**
 151 **structure and flammability.**

Leaf trait	Definition	Effect on litter bed structure*	Association with litter flammability*	Studies observing relationship
Length (mm)	Length of leaf	bulk density (-)	combustibility (+) consumability (+)	(de Magalhães and Schwilk, 2012a, Scarff and Westoby, 2006, Engber and Varner, 2012, Schwilk and Caprio, 2011)
Leaf curl (mm)	Maximum absolute height of leaf when positioned on a flat surface	bulk density (-) packing ratio (-)	combustibility (+) sustainability (-) consumability (+)	(Engber and Varner, 2012, Grootemaat et al., 2017a)
Surface area to volume ratio (SAV, cm ⁻¹)	Two-sided leaf area per leaf volume	packing ratio (-)	ignitability (-) consumability (+)	(Rothermel, 1972, Parsons et al., 2015)
Specific leaf area (SLA, cm ² g ⁻¹)	One-sided leaf area per leaf dry mass	bulk density (-)	ignitability (-) sustainability (-)	(Grootemaat et al., 2017a, Parsons et al., 2015, Ganteaume, 2018, de Magalhães and Schwilk, 2012a, Murray et al., 2013)
<i>Inverse is leaf mass per area (LMA, g cm⁻²)</i>			combustibility (+)	
Area (cm ²)	One-sided leaf area	packing ratio (-) bulk density (-)	combustibility (+) sustainability (-)	(Scarff and Westoby, 2006, de Magalhães and Schwilk, 2012a, Cornwell et al., 2015a, Murray et al., 2013)

Thickness (mm)	Thickness of leaf at intermediate position between border and midrib	packing ratio (+)	ignitability (+)	(Parsons et al., 2015, Plucinski and Anderson, 2008, Ganteaume, 2018)
Tissue density (g cm ⁻³)	Leaf dry mass per leaf dry volume	bulk density (+)	sustainability (+)	(van Altena et al., 2012, Grootemaat et al., 2017a)
			sustainability (+)	
			combustibility (-)	

* + or – indicate the direction of influence of leaf trait on litter structure or flammability metric, e.g. leaf length has a negative influence on bulk density, longer leaves create lower bulk density litter beds.

152 2. Materials and methods

153 2.1. Literature review and data compilation

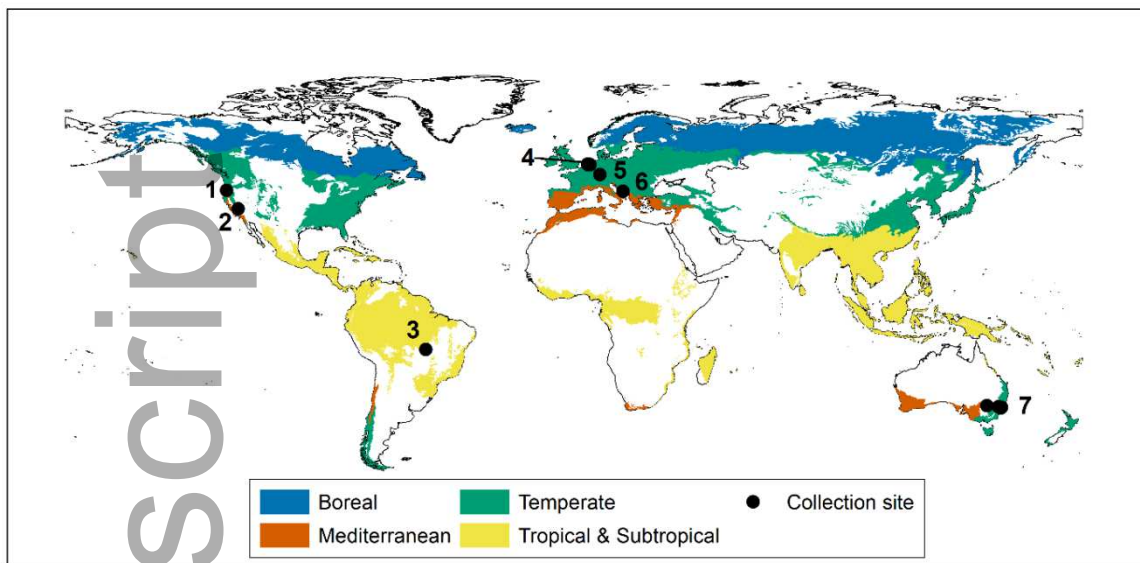
154 A structured literature review was conducted using the Web of Science database (all databases).
 155 Document types were restricted to published journal articles and the time span included all
 156 years (1900-2019). Keywords required were “litter”, “flammability” and “traits”. This search
 157 yielded 55 potential articles for the analysis. From an initial screening of the literature, the
 158 following criteria were developed and then applied to select appropriate studies to include in
 159 the analysis:

- 160 1. laboratory studies testing the flammability of single-species leaf litter beds (not
 161 individual leaves);
- 162 2. measured at least two of three flammability components: sustainability (flame
 163 duration, seconds), combustibility (flame spread rate, cm s⁻¹) and consumability
 164 (consumption, % mass consumed), ignitability was not included as it was not
 165 measured in a consistent way across studies;
- 166 3. litter structure was measured as either litter depth, bulk density or packing ratio;
 167 and
- 168 4. at least one leaf trait was measured.

169 Seven studies were selected based on the above criteria, creating a final dataset that consisted
 170 of 109 observations from 106 unique woody species across four forest biomes (Fig. 1, Table

171 S1, Table S2). Africa and Asia were not well-represented in the dataset as no comparable
172 studies have been conducted in these regions. All studies contained multiple species, ranging
173 from three (Kauf et al., 2019) to 33 species (Cornwell et al., 2015a). Three plant species were
174 measured across multiple studies (*Calocedrus decurrens*, *Quercus kelloggii* and
175 *Sequoiadendron giganteum*). The Cornwell et al. (2015a) dataset comprised conifer litter
176 from different regions including Europe, Asia, South Africa, South America, North America
177 and Australia and was collected from several temperate arboretums, botanic gardens and
178 greenhouses in the Netherlands. Similarly, the Kauf et al. (2018) comprised litter collected
179 from botanic gardens in Germany. Although environmental conditions in the botanic gardens
180 may differ from the native environment of a given species, in this study, we were interested
181 in general trait-flammability relationships, so the use of material from botanic gardens is
182 justified. See Table S2 for the full list of plant species across all studies. Data were extracted
183 from online repositories for most studies (de Magalhães, and Schwilk, 2012b, Cornwell et al.,
184 2015b, Parsons and Balch, 2015a, Parsons and Balch, 2015b, Grootemaat et al., 2017b). For
185 the remaining studies, data were extracted from the main text, tables and/or digitised graphs
186 of primary references. Where data were collected but not reported, the authors were contacted
187 and data incorporated if they were made available. All experiments were conducted on dry (<
188 10% moisture content) litter beds. Details of experimental set-up and flammability metrics
189 are in Table S1. Plant species litter was categorised into four groups based on litter type: leaf
190 (flattened leaf, includes both conifer and broad-leaf species), short-needle, long-needle and
191 branch. Branch litter included coniferous species from Cupressaceae and Araucariaceae
192 which typically do not shed individual leaves but small branches with leaves attached.
193 However, for simplicity, we use the term ‘leaf’ to describe any of the litter types.

194 Litter structure was measured as packing ratio and bulk density. In cases where bulk density
195 was not reported it was calculated from sample mass and fuel bed volume based on litter
196 depth and sample tray area. This was necessary for species reported in Engber and Varner
197 (2012) and Parsons et al. (2015). Where packing ratio was not recorded it was calculated by
198 dividing bulk density by particle density as was done for the Parsons et al. (2015) dataset.
199 Supplementary leaf trait data for the species from the Cornwell et al. (2015a) dataset were
200 sourced from the TRY database (Kattge et al., 2020) and The Gymnosperm Database (Earle,
201 2020).



203 **Figure 1.** Location of studies included in the meta-analysis. Colours represent different forest
 204 biomes (from WWF Terrestrial Ecoregions <https://www.worldwildlife.org>). Black circles
 205 indicate collection sites. Studies are as follows: 1 Engber and Varner, 2012; 2 de Magalhães
 206 and Schwilk, 2012a; 3 Parsons et al., 2015; 4 Cornwell et al., 2015a, Zhao et al., 2016; 5
 207 Kauf et al., 2018; 6 Kauf et al., 2019 and 7 Grootemaat et al., 2017a. Note Study 4 includes
 208 species from multiple biomes but collected from temperate aboretums, botanic gardens and
 209 subtropical greenhouses.

210 2.2. Data analysis

211 All leaf trait, litter structure data and flammability metrics were log-transformed to meet
 212 assumptions of normality and homogeneity of residuals prior to analysis. The only exception
 213 was consumption for which we used raw values for the analysis because a non-normal
 214 distribution was observed.

215 We first tested how leaf traits influence both litter structure and flammability. Seven leaf
 216 traits were identified as important based on previous studies (Table 1). Initially, we used
 217 Pearson correlation to test for relationships between the seven leaf traits, litter structure and
 218 flammability (Fig. S1). From this, we selected key traits which had a strong ($r > 0.7$)
 219 correlation to either litter structure or flammability. This led us to choose three leaf traits for
 220 inclusion in further analysis; leaf curl, SAV and SLA. Relationships between selected leaf
 221 traits, litter structure and flammability metrics were then assessed using linear regression with
 222 a gaussian error distribution. For the relationship between packing ratio and consumption we
 223 used logistic regression with a binomial distribution, following Cornwell et al. (2015a) with

224 consumption treated as a proportion. We evaluated the strength of linear regressions using R^2
225 values and for the logistic regression we used log likelihood. All bivariate analyses were
226 performed in R version 3.6.2 (R-Core-Team, 2016).

227 Piecewise Structural Equation Modelling (SEM) was used to disentangle the effects of
228 different leaf traits and litter structure on litter flammability. SEM can incorporate multiple
229 predictors and response variables in a single causal network, allowing direct and indirect
230 effect pathways to be evaluated (Grace, 2006, Grace et al., 2010). Piecewise SEM is an
231 extension of the traditional SEM approach. Here, the SEM is translated into a set of linear
232 models which are evaluated individually or locally and then combined to make inferences
233 about the entire SEM (Lefcheck, 2016). This allows the source of the data to be included as a
234 random effect to account for the effect of different experimental approaches and conditions.
235 Model conceptualisation and parametrisation was based on a priori knowledge and our
236 bivariate analysis. Leaf traits which had a moderately strong relationship ($R^2 \geq 0.50$) with
237 litter structure or flammability were chosen as predictors. Thus, SLA and leaf curl were
238 chosen as predictors of bulk density. SAV and leaf curl were chosen as predictors of packing
239 ratio. Leaf curl and SLA were chosen as predictors of flame duration. SLA was chosen as a
240 predictor of flame spread rate. For flame spread rate and flame duration, both packing ratio
241 and bulk density were chosen as predictors, in addition to selected leaf traits. Only packing
242 ratio was selected for consumption with no leaf traits selected as predictors due to the lack of
243 strong associations.

244 Each component model of the piecewise SEM was set-up as a linear mixed effect model
245 (LMM), with errors estimated using restricted maximum likelihood and study number
246 included as a random effect. As the variables measured differed between studies, the number
247 of observations and studies included in each LMM differed from 44 observations (4 studies)
248 to 83 observations (5 studies). Variables were scaled and centred prior to analysis. We report
249 the standardised path coefficient for each path, which indicates the relative strength of the
250 different predictors on litter structure or flammability. We also report two different R^2 for
251 each response variable, the marginal R^2 (R_m^2 , based on the variance of the fixed effects only)
252 and the conditional R^2 (R_c^2 , based on the variance of both the fixed and random effects). We
253 evaluated the fit of the entire SEM using the Fisher's C test, with small Fisher's C and $p >$
254 0.05 indicating good model fit. Piecewise SEM was performed in R version 3.6.2 (R-Core-
255 Team, 2016), using the packages 'lme4' (Bates et al., 2015) and 'piecewiseSEM' (Lefcheck,
256 2016). Code and data are available via Figshare (Burton et al., 2020).

257 3. Results

258 3.1. Interspecific differences in litter structure and flammability

259 Litter structure and flammability varied among the 106 shrub and tree species in the dataset
260 (Table 2, Fig. 2). There was little variation in consumption for most species, short-needle
261 conifers had the lowest amount of material burnt (< 10% consumed). Flame spread rate and
262 flame duration were significantly negatively correlated ($r = -0.78$, $p < 0.001$, Fig. S1),
263 meaning that when flames spread more slowly through the litter bed, the burning was
264 sustained for longer. Consumption was not significantly correlated with either flame spread
265 rate nor flame duration (Fig. S1).

266 **Table 2.** Summary statistics for litter structure and flammability components.

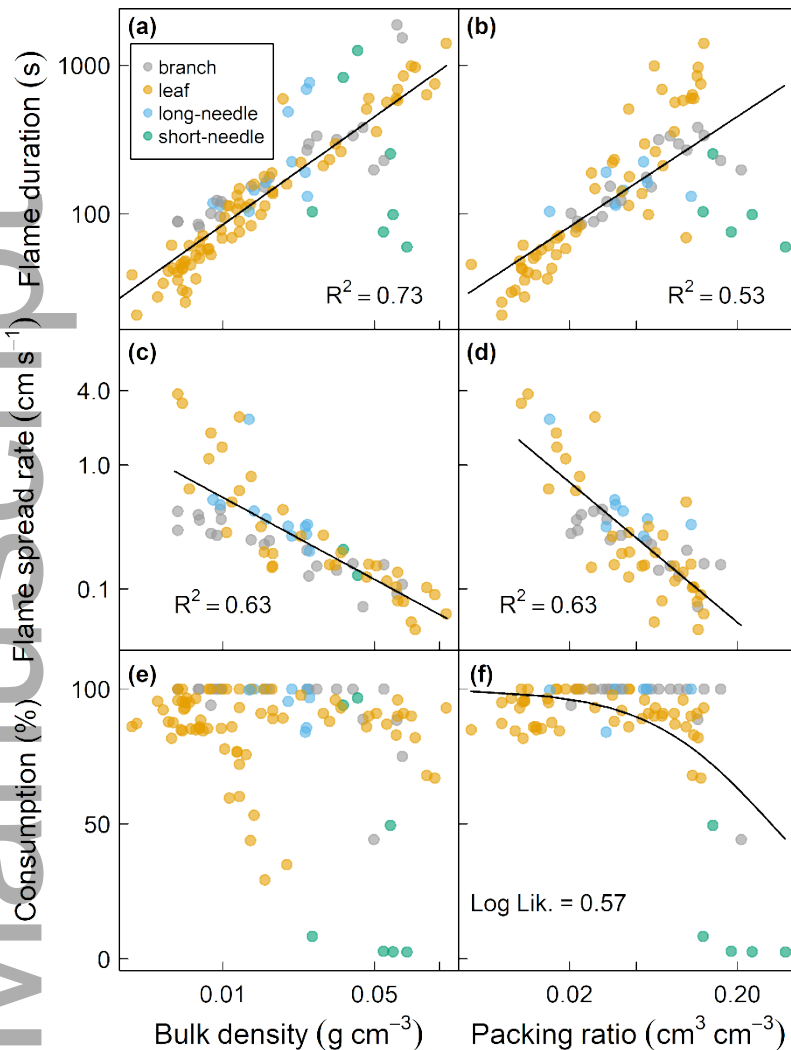
Variable	Mean (min – max)	x-fold variance
<i>Litter structure</i>		
Packing ratio ($\text{cm}^3 \text{cm}^{-3}$)	0.06 (0.005 – 0.39)	74.3
Bulk density (g cm^{-3})	0.02 (0.003 – 0.10)	28.0
<i>Flammability</i>		
Sustainability, flame duration (s)	258 (21 – 1880)	90.5
Combustibility, flame spread rate (cm s^{-1})	0.5 (0.05 – 3.7)	79.7
Consumability, consumption (%)	85 (3 – 100)	38.1

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268 Flame duration was better predicted by bulk density (slope = 1.04, $R^2 = 0.73$, $p < 0.05$) than
269 packing ratio (slope = 0.74, $R^2 = 0.53$, $p < 0.05$) (Fig. 2a, 2b). Litter beds that were more
270 aerated (lower packing ratios and lower bulk densities) had shorter flame durations. Flame
271 spread rate declined with increasing bulk density (slope = -0.95, $R^2 = 0.63$, $p < 0.05$) and
272 packing ratio (slope = -1.13, $R^2 = 0.63$, $p < 0.05$) (Fig. 2c, 2d). There was no relationship
273 between bulk density and consumption (Fig. 2e). However, packing ratio had a significant
274 relationship with consumption (slope = -1.17, $p = 0.01$, log likelihood = 0.57). Consumption
275 decreased with increasing packing ratio, with substantially lower consumption observed
276 beyond $0.15 \text{ cm}^3 \text{cm}^{-3}$ (Fig. 2f). Litter derived from short-needle coniferous plant species
277 (*Picea* sp., *Abies* sp., *Tsuga* sp. and *Larix* sp.) did not conform to the general trends observed
278 for other litter types. Short-needle litter were characterised by higher bulk densities and
279 packing ratios and lower litter consumption.

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Figure 2.



293 Relationships between litter structure and flammability from experimental burns. Each dot
294 represents a species mean. Significant ($p < 0.001$) relationships from linear (a-d) and logistic
295 (f) regression are represented with solid lines and corresponding R^2 (linear) and Log
296 Likelihood values (Log Lik, logistic regression). All axes are log-transformed except for
297 consumption.

298 3.2. Which leaf traits are most important for litter structure and flammability?

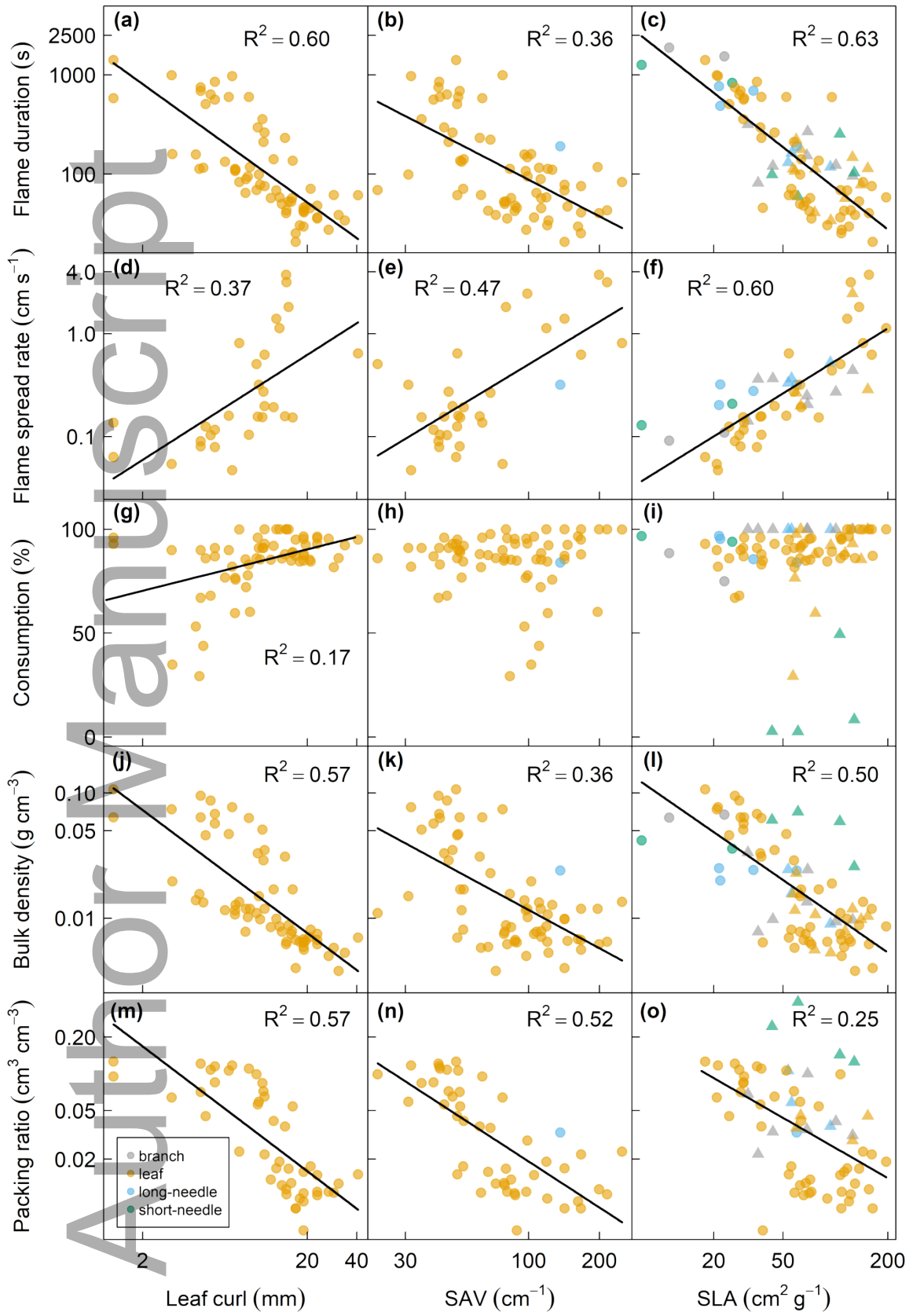
299 Leaf area, length, thickness and density were not strongly related to litter bed structure or
300 flammability (Fig. S1). In contrast, leaf curl, SLA and SAV showed stronger associations
301 with litter bed structure and some flammability metrics (Fig. S1).

302 The strongest predictors of flame duration were SLA (slope = -1.38, $R^2 = 0.63$, $p < 0.05$) and
303 leaf curl (slope = -1.19, $R^2 = 0.60$, $p < 0.05$) (Fig. 3a, 3c). Leaves with higher SLA and
304 greater leaf curl had shorter burning times than leaves that had lower SLA and leaf curl. SLA

305 was also the best predictor of flame spread rate, explaining 60% of the variation in flame
306 spread rate (slope = 1.05, $R^2 = 0.60$, $p < 0.05$) (Fig. 3f). Litter beds composed of leaves with
307 greater SLA had faster flame spread rates. Consumption did not have a strong relationship to
308 any of the leaf traits, being only weakly positively related to leaf curl (Fig. 3g-i).

309 Bulk density was negatively related to SLA, leaf curl and SAV (Fig. 3j-l). Leaf curl (slope = -
310 0.98, $R^2 = 0.57$, $p < 0.05$) and SLA (slope = -0.96, $R^2 = 0.50$, $p < 0.05$) were the strongest
311 drivers of bulk density, explaining 57% and 50% of the variation in bulk density respectively.
312 Packing ratio exhibited similar trends to leaf traits as bulk density (Fig. 3m-o). However, leaf
313 curl (slope = -1.02, $R^2 = 0.57$, $p < 0.05$) and SAV (slope = -1.25, $R^2 = 0.52$, $p < 0.05$) were
314 the strongest drivers of packing ratio, explaining 57% and 52% of the variation in packing
315 ratio respectively.

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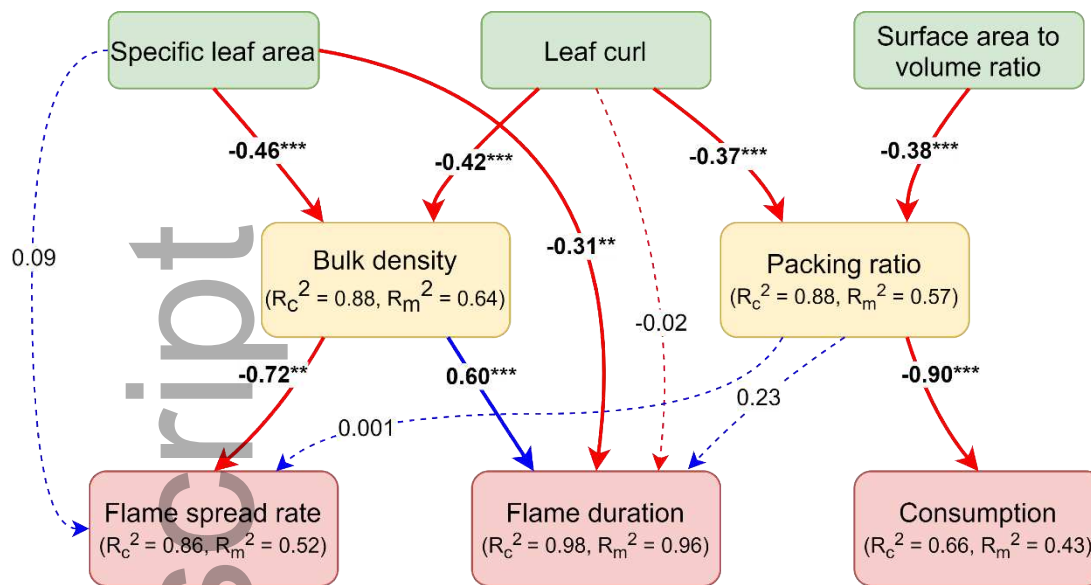
317 **Figure 3.** Relationship between leaf traits, litter structure and flammability from
318 experimental burns. Each symbol represents a species mean. Shapes denote whether leaf trait
319 (SLA) was measured within the study (circle) or sourced from an alternative source
320 (triangle). Significant ($p < 0.001$) relationships are represented with solid lines and
321 corresponding R^2 values. All axes are log-transformed except for consumption. SAV =
322 surface area to volume ratio, SLA = specific leaf area.

323 *3.3. Influence of leaf traits and litter structure on flammability*

324 The Piecewise Structural Equation Model was a good fit to the data (Fisher's $C = 30.87$, $p =$
325 0.23 , $df = 26$). There was a difference between the conditional R^2 and marginal R^2 for flame
326 spread rate and consumption but not flame duration. Hereafter, we refer to the marginal R^2
327 only as this represents the effects of predictor variables only. Leaf curl and SAV explained
328 57% of the variation in packing ratio (Fig. 4). SAV and leaf curl both had a negative effect on
329 packing ratio, with similar path coefficients (SAV, coefficient = -0.38 , $p < 0.001$ and curl,
330 coefficient = -0.37 , $p < 0.001$). Leaf curl and SLA explained 64% of the variation in bulk
331 density, with SLA and curl both having a negative effect (coefficient = -0.46 , $p < 0.001$ vs
332 coefficient = -0.42 , $p < 0.001$).

333 Leaf curl did not have a significant influence on flame duration despite being strongly related
334 in bivariate analysis (coefficient = -0.02 , $p = 0.65$). SLA did have a significant negative effect
335 on flame duration (coefficient = -0.31 , $p = 0.003$) but not flame spread rate (coefficient =
336 0.09 , $p = 0.59$). Bulk density significantly positively influenced flame duration (coefficient =
337 0.60 , $p < 0.001$) and negatively influenced flame spread rate (coefficient = -0.72 , $p = 0.003$).
338 Packing ratio had a significant negative effect consumption (coefficient = -0.90 , $p < 0.001$)
339 but was not significantly related to flame duration (coefficient = 0.23 , $p = 0.06$) nor flame
340 spread rate (coefficient = 0.001 , $p = 0.99$) in the piecewise SEM. Packing ratio explained
341 43% of the variation in consumption. The combination of leaf traits and litter structure
342 explained 52% and 96% of the variation in flame spread rate and flame duration respectively.

343



344

345 **Figure 4.** Piecewise Structural Equation Model for the effect of leaf traits on litter structure
 346 and flammability. Solid arrows indicate significant relationships between variables, and
 347 dashed arrows indicate non-significant relationships. Blue arrows indicate positive
 348 relationships and red arrows indicate negative relationships. Numbers on arrows show
 349 standardised path coefficients. Significant values are indicated by * for $p < 0.05$, ** for $p <$
 350 0.01 and *** for $p < 0.001$. Conditional R^2 and marginal R^2 values shown in the boxes of the
 351 response variables.

352 4. Discussion

353 4.1. Importance of leaf traits to litter bed structure and flammability

354 We quantified relationships between leaf traits, litter structure and flammability for a wide
355 range of plant species (and litter types) across multiple forest biomes. Despite some
356 differences in experimental design, we found consistent significant effects of leaf traits on
357 litter structure and flammability. Specifically, we found traits which characterise the three-
358 dimensional nature of the leaf and how much space a leaf occupies (namely leaf curl, SAV
359 and SLA) showed much stronger associations to litter structure and flammability than other
360 morphological traits (thickness, length, area and tissue density).

361 Leaf curl influences both packing ratio and bulk density. Leaves with greater curl allow more
362 air spaces in the litter bed leading to lower packing ratios and bulk densities (Engber and
363 Varner, 2012, Grootemaat et al., 2017a). In addition to leaf curl, SAV drove variation in
364 packing ratio, with larger leaves resulting in more aerated litter (lower packing ratio), as
365 previously found (Grootemaat et al., 2017a). Leaf curl and SLA were strongly associated
366 with bulk density. Leaves which were less curled and had lower SLA created denser litter
367 beds (higher bulk density). Leaf curl covaries with leaf length and size such that larger leaves
368 also have greater curl. Thus, it is possible that the relationship between curl and litter bed
369 packing may be related to differences in leaf length and size. Isolating the effect of
370 leaf curl would require experimental manipulations standardising curl relative to leaf size. In
371 addition, leaf curl data were available for leaf litter beds only, not needle or branch litter. For
372 needle and branch litter, litter particle size (measured as litter particle length \times width \times
373 height) which also indicates the three-dimensional nature of the litter particle, may be a better
374 determinant of packing ratio (Zhao et al., 2016).

375 Morphological leaf traits were not related to consumption in contrast to previous studies that
376 found consumption to be positively related to both SAV (Parsons et al., 2015) and
377 community-weighted mean leaf length (Schwilk and Caprio, 2011). It is possible that traits
378 related to chemical composition rather than morphology may be more important for driving
379 differences in consumption. Grootemaat et al. (2017a) found that for sclerophyllous south-
380 eastern Australian species, higher tannin and lower lignin concentrations were associated
381 with lower consumption. Insufficient data about the chemical traits across the range of plant
382 species in our study meant we could not test for the effect of these on flammability, making it
383 a worthwhile avenue for further exploration.

384 Morphological leaf traits were strongly associated with flame spread rate and duration,
385 consistent with previous studies (Engber and Varner, 2012, Grootemaat et al., 2017a, de

386 Magalhães and Schwilk, 2012a). Flame duration was most strongly negatively related to leaf
387 curl and SLA. Litter composed of curlier particles with greater SLA flamed for shorter
388 periods. Flame spread rate was strongly positively related to SLA. Faster flame spread rates
389 were observed in litter composed of leaves with greater SLA. However, due to the
390 interrelatedness between leaf traits, litter structure and flammability, it is not possible to
391 determine whether these associations are from a direct effect or indirect effect via litter
392 structure or a combination of both. By using piecewise SEM, we were able to quantify the
393 direct and indirect effect (via litter structure) of leaf traits on litter bed flammability, the
394 results of which are discussed below.

395 *4.2. Direct and indirect effects of leaf traits on flammability*

396 Litter bed flammability is driven largely by litter structure. Flame duration was the only
397 flammability metric to be influenced by both leaf traits and litter structure in the SEM, being
398 negatively related to SLA and positively related to bulk density. The direct effect of SLA on
399 flame duration is likely related to the influence of SLA on leaf ignitability. Leaves with
400 higher SLA have been found to ignite more quickly owing to the greater surface area exposed
401 to heat and flames (Grootemaat et al., 2015, Murray et al., 2013, Ganteaume, 2018). The
402 spread of fire in the litter depends on the successive ignitions of individual leaves
403 (Grootemaat et al., 2017a, Martin et al., 1994). Thus, the more easily leaves ignite, the faster
404 flames can spread from leaf to leaf, leading to shorter flame durations overall.

405 Bulk density emerged as the most important predictor of flame spread rate and flame
406 duration. Litter beds with higher bulk density have more fuel available to burn but longer
407 flame durations and slower flame spread rates. This is likely due to variation in the
408 penetration distance of radiant heat. At higher bulk densities radiative heat transfer through
409 the fuel bed is lower, leading to longer burning times and slower spread rates (Bartoli et al.
410 2011). Unlike flame duration, flame spread rate was not significantly related to any leaf traits
411 in the SEM. In addition, leaf curl did not have a significant effect on flame duration, despite
412 being strongly negatively related in the bivariate analysis. Thus, we suggest the strong
413 associations between SLA and flame spread rate and leaf curl and flame duration in bivariate
414 analysis is an artefact of their influence on bulk density rather than a direct effect on
415 flammability.

416 Packing ratio is a strong driver of consumption, explaining 43% of the variation in
417 consumption. At low packing ratios air can easily flow into the litter bed providing well-
418 ventilated combustion conditions, which results in greater consumption as downwards heat
419 transfer is not limited. At higher packing ratios and no-wind conditions, combustion becomes

420 increasingly limited in oxygen supply (El Houssami et al., 2018). This results in lower
421 amounts of material consumed, as observed in our dataset. There was a sharp decline in
422 consumption at higher packing ratios ($0.15 \text{ cm}^3 \text{ cm}^{-3}$), which is consistent with previous
423 studies (Cornwell et al., 2015a, Zhao et al., 2016). This decline was driven by the presence of
424 a few short-needle conifers (*Larix* sp., *Abies* sp., *Tsuga* sp. and *Picea* sp.), making it difficult
425 to confirm whether the patterns observed are due to packing ratio alone or some other
426 unquantified trait potentially related to litter chemistry as previously outlined (see section
427 4.1.). However, there is qualitative evidence of broad-leaf litter, similar in size and shape to
428 the short-needle coniferous species, failing to ignite at high packing ratios. Plucinski and
429 Catchpole (2001) who used the same methods as (Cornwell et al., 2015a, Grootemaat et al.,
430 2017a) reported that litter from *Banksia ericifolia* L.f. (packing ratio = 0.284, leaf area =
431 0.242 cm^2) and *Leptospermum laevigatum* (Gaertn.) F.Muell. (packing ratio = 0.202, leaf
432 area = 0.576 cm^2) failed to ignite and spread at low moisture contents. This suggests that the
433 packing ratio-consumption relationship may be relevant to both broad-leaf and conifer
434 species.

435 It is important to note that all experiments in our meta-analysis were conducted in controlled
436 conditions with no effect of wind or radiant heat. It would be interesting to explore under what
437 ambient conditions ignition is successful and if the effect of packing ratio becomes less
438 influential. For example, the presence of wind may support ignition and burning rate by
439 supplying oxygen, mixing pyrolysis gases with air (El Houssami et al., 2018) and increasing
440 flame contact from the ignition source to the fuel (Rothermel, 1972, Plucinski and Catchpole,
441 2001). Another important consideration is the influence of variation in experimental design on
442 the results. It is possible that difference in burning methods, such as tray shape (e.g. square,
443 rectangle and circle) and ignition source contributed to some of the variation observed in flame
444 spread rate and consumption. To this end, it is important that any future studies follow
445 consistent methods, perhaps following methods in Cornwell et al. (2015a) as this methodology
446 represented half of the dataset in our study (56 out of 109 observations).

447 4.3. Applications and next steps

448 Our study provides insight into how flammability and the traits which drive flammability
449 differ across taxa. Two genera which occur in fire-prone environments, *Eucalyptus* and
450 *Pinus*, create litter with low packing ratios and high flammability. However, they produce
451 contrasting litter particles, large curled leaves vs. long thin needles grouped in fascicles. In
452 comparison the non-*Pinus* Pinaceae create litter which is distinct from any of the other
453 species in the dataset as highlighted by Cornwell et al. (2015a). Future studies could integrate
454 the flammability data synthesised in this study with species phylogeny to explore the

455 evolutionary patterns of litter bed flammability, similar to the approach of Cui et al. (2020)
456 for shoot flammability and expanding on the analysis of Zhao et al. (2016).

457 Identifying the underlying traits responsible for variation in flammability can allow the
458 prediction of flammability using species-specific traits. Although there are models for live
459 plants (Forest Flammability Model, Zylstra et al., 2016), there is currently no equivalent
460 model for litter beds. This study represents a significant step towards developing trait-based
461 empirical models for litter beds. Such models could be incorporated into existing models
462 which use plant traits (e.g. Forest Flammability Model) to improve the estimation of surface
463 fire behaviour. Trait-based models could also be developed to predict the ecological effects of
464 wildfires. For example, flame spread rate in addition to flame duration, influences soil
465 temperatures and thus soil heating. The degree of soil heating influences physical soil
466 properties (Certini, 2005), the mortality of soil organisms (Pingree and Kobziar, 2019) and
467 the germination of buried dormant seeds (Bradstock et al., 1992, Burrows, 1999, Penman and
468 Towerton, 2008). In addition, flame spread rate and fuel consumption are sub-components of
469 fire intensity (Bryam, 1959) which is important for understanding wildfire effects above-
470 ground. Fire intensity and flame duration influence plant mortality, with longer flame
471 durations and higher fire intensities causing greater damage (Neary et al., 1999, Gagnon et
472 al., 2010).

473 Trait-based models require information on both senesced and green leaf traits, as leaf traits
474 most important to flammability in a living plant differ from those most important in a litter
475 bed. However, the collection of trait data to understand flammability is heavily skewed
476 towards traits that are important for living vegetation (Pérez-Harguindeguy et al., 2013). SLA
477 is an important trait for the ignitability of green leaves (Grootemaat et al., 2015, Murray et
478 al., 2013) and here we show it affects flammability directly (flame duration) and indirectly
479 via litter structure (bulk density). However, green leaf SLA may not be the same as senesced
480 leaf SLA due to carbon and nutrient resorption before abscission (Vergutz et al., 2012). Thus,
481 there is a need to measure SLA of both green and freshly senesced leaves. Another trait for
482 which there is no protocol for measurement in the handbook for leaf traits is leaf curl. As the
483 value of leaf curl may change with moisture or decomposition stage, measurements should be
484 taken from dried freshly senesced, undecomposed leaves. This would also reflect the
485 maximum potential leaf curl of a species. An interesting avenue for future research could be
486 exploring what influences the tendency for leaves to curl when senesced. If curl can be
487 associated with green leaf traits, predictions of curl (and thus packing ratio) could then be
488 made using green leaf traits for which there is greater data availability.

489 Another important consideration is the temporal variability of leaf trait effects on litter bed
490 flammability. The relationships we identified are most relevant for freshly senesced dry (<
491 10% moisture content) leaf litter. However, leaf traits and litter beds are not static and are
492 continually changing over time. Thus, the influence of these traits on flammability may
493 fluctuate due to changes in moisture, decomposition and fragmentation. Exposure to moisture
494 can reduce leaf curl and increase compaction of the litter bed, which increases bulk density
495 (Kauf et al., 2018, Weir and Limb, 2013). Decomposition and fragmentation affect the size,
496 shape and chemical composition of the leaves in the litter bed which changes litter structure
497 and flammability (Cornwell et al., 2009, Kauf et al., 2015, Zhao et al., 2014). The temporal
498 variability of leaf trait effects on flammability is an interesting question for future studies.

499 The amount and composition of the litter bed are also important determinants of litter
500 flammability, in addition to leaf traits and litter structure (Varner et al., 2015). Thus, any
501 possible trait-based model of litter flammability needs to include these effects. The amount of
502 litter depends on the rate of litterfall and decomposition which varies with site productivity
503 and climate (Facelli and Pickett, 1991). In natural forests litter beds are rarely homogenous
504 and are usually composed of more than one species or component (e.g. leaf, twig, bark).
505 Several studies have investigated the effect of compositional variation by sampling intact
506 samples (Ganteaume et al., 2014) or by artificially creating leaf litter mixtures of two or more
507 species (de Magalhães and Schwilk, 2012a, Della Rocca et al., 2018, van Altena et al., 2012,
508 Zhao et al., 2016). There is growing evidence that litter mixtures can behave non-additively,
509 that is the flammability of the mixture differs markedly from expected flammability based on
510 the flammability of the species when burnt in isolation. Moreover, moisture content can
511 influence the magnitude and direction of non-additivity (Blauw et al., 2015, de Magalhães
512 and Schwilk, 2020). However, the majority of these studies have used mixtures made up of
513 the same litter component (leaves, needles) from different plant species, with fewer studies
514 using mixtures made up of different litter components e.g. leaves, bark and twigs (but see
515 (Zhao et al., 2019, Gormley et al., 2020)). Further research into the flammability of mixed
516 litter beds (species and components) is required to better understand litter bed flammability.

517 We show that trait-flammability relationships previously identified in individual studies are
518 maintained for a wide-range of plant species across multiple forest biomes. Our synthesis
519 represents a crucial first step towards developing empirical models of litter bed flammability
520 using leaf traits. Such models will be useful for understanding flammability in plant
521 communities now and into the future as altered fire regimes and climate change trigger shifts
522 in plant species composition and leaf traits.

523 **Acknowledgements**

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524 JEB was funded by an Australian Government Research Training Program (RTP)
525 Scholarship. Funding for JGC is from the Victorian Department of Environment, Land,
526 Water and Planning. AIF is funded by the Bushfire and Natural Hazards Cooperative
527 Research Centre. We thank the three anonymous reviewers and associate editor for their
528 constructive comments which greatly improved the quality of the manuscript. We declare that
529 the authors do not have any conflict of interest.

530 **Author's contributions**

531 JEB conceptualised the meta-analysis. JEB designed the methodology, collated the data and
532 performed the analysis. JEB led the writing of the manuscript; TDP, JGC and AIF are
533 supervisors of JEB PhD project and as such made substantial contributions to project design
534 and the writing and revision of the manuscript. They approved the final version for
535 publication.

536 **Data availability statement**

537 Data and code associated with this manuscript are archived and made publicly available on
538 Figshare: <https://doi.org/10.6084/m9.figshare.13218836.v3> (Burton et al., 2020).

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