

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25

**Sex in troubled waters: widespread agricultural contaminant disrupts
reproductive behaviour in fish**

Michael G. Bertram ^{a,*}, Minna Saaristo ^{a,b}, John B. Baumgartner ^c, Christopher P.
Johnstone ^a, Mayumi Allinson ^d, Graeme Allinson ^{d,e}, Bob B.M. Wong ^a

^a *School of Biological Sciences, Monash University, Victoria, Australia*

^b *Department of Biosciences, Åbo Akademi University, Turku, Finland*

^c *ARC Centre of Excellence for Environmental Decisions, School of Botany, The
University of Melbourne, Victoria, Australia*

^d *Centre for Aquatic Pollution Identification and Management (CAPIM), The
University of Melbourne, Bio21 Institute, Victoria, Australia*

^e *Department of Environment and Primary Industries (DEPI), Victoria, Australia*

* Corresponding author at:

School of Biological Sciences, Monash University, Senior Zoology Bldg. 18,
Wellington Road, Clayton Campus, Victoria 3800, Australia

E-mail address: michael.g.bertram@monash.edu (M.G. Bertram); +61 3 9902 0756

26 **Abstract**

27 Chemical pollution is a pervasive and insidious agent of environmental change. One
28 class of chemical pollutant threatening ecosystems globally are the endocrine
29 disrupting chemicals (EDCs). The capacity of EDCs to disrupt development and
30 reproduction is well established, but their effects on behaviour have received far less
31 attention. Here, we investigate the impact of a widespread androgenic EDC on
32 reproductive behaviour in the guppy, *Poecilia reticulata*. We found that short-term
33 exposure to an environmentally relevant concentration of 17 β -trenbolone — a
34 common environmental pollutant associated with livestock production — influenced
35 the amount of male courtship, and forced copulatory behaviour (sneaking), performed
36 toward females. Exposure to 17 β -trenbolone was also associated with greater male
37 mass. However, no effect of exposure was detected in females, indicating sex-specific
38 vulnerability at this dosage. Our study is the first to show altered male reproductive
39 behaviour following exposure to an environmentally realistic concentration of 17 β -
40 trenbolone, demonstrating the possibility of widespread disruption of mating systems
41 of aquatic organisms by common agricultural contaminants.

42

43 *Keywords*

44 Endocrine disrupting chemical, EDC, hormonal growth promotant, 17 β -trenbolone,
45 trenbolone acetate, guppy, *Poecilia reticulata*, behavioural ecotoxicology, sexual
46 selection, reproductive behaviour.

47

48

49

50

51 **Introduction**

52 Chemical pollutants have accumulated in ecosystems globally, endangering wildlife,
53 ecosystem function and human health (Schwarzenbach et al., 2006). One class of
54 chemical pollutant, known as endocrine disrupting chemicals (EDCs), comprises
55 environmental contaminants with the capacity to disrupt the natural hormonal
56 functioning of organisms (Colborn et al., 1993). Endocrine disruptors are of particular
57 concern given their extreme potency, with exposure to concentrations as low as
58 nanograms per litre having deleterious effects, as well as the propensity of some
59 EDCs to bioaccumulate, persist temporally and act transgenerationally (Diamanti-
60 Kandarakis et al., 2009). Conventionally, studies in ecotoxicology have focussed on
61 direct mortality and chronic sub-lethal effects of EDCs on development and
62 reproduction (Melvin and Wilson, 2013). However, EDCs can also induce alarming
63 changes in behaviour. Indeed, the particular sensitivity of behaviour to EDCs has
64 driven recent interest in behavioural ecotoxicology as a tool for investigating
65 endocrine disruption at environmentally relevant pollutant concentrations (reviewed
66 in Melvin and Wilson, 2013). Existing studies in behavioural ecotoxicology typically
67 focus on EDCs that disrupt gonadal steroid signalling by interacting with vertebrate
68 estrogen or androgen receptors, as chemical interference with this pathway has the
69 potential to disrupt sexual selection (e.g., Saaristo et al., 2009). However, the vast
70 majority of these efforts have concentrated on EDCs with estrogenic activity. This is
71 surprising because the handful of studies that have considered androgenic EDCs
72 suggest that they are also capable of markedly altering animal behaviour (e.g.,
73 Hoffmann and Kloas, 2012).

74 An androgenic EDC of particular concern is 17 β -trenbolone, the most
75 bioactive metabolite of trenbolone acetate, a hormonal growth promotant used

76 extensively in livestock production around the world (Kolodziej et al., 2013).
77 Trenbolone acetate is a powerful steroid, with androgenic and anabolic potency 15–50
78 times greater than testosterone (Kolodziej et al., 2013; Neumann, 1976). Its
79 metabolite 17 β -trenbolone acts as a powerful androgen agonist in the environment, is
80 highly temporally persistent (with a half-life of approximately 260 days; Schiffer et
81 al., 2001) and has been repeatedly detected in aquatic environments associated with
82 feedlot operations. Detected concentrations of 17 β -trenbolone range from \leq 20 ng/L in
83 diffuse run-off (Durhan et al., 2006), to as high as 162 ng/L in fields directly receiving
84 effluent (Gall et al., 2011). Recent studies report that exposure to 17 β -trenbolone
85 adversely impacts physiological and morphological endpoints in fish species (e.g.,
86 Morthorst et al., 2010). However, despite the potency and widespread global use of
87 17 β -trenbolone, very little is known about its effects on behaviour. This is concerning
88 as the ability of animals to produce and maintain behaviour appropriate to their
89 environment is fundamental for survival and reproduction, so that disruption of these
90 behaviours can have dire ecological and evolutionary consequences (reviewed in
91 Candolin and Wong, 2012).

92 The mating system of the guppy, *Poecilia reticulata*, is ideal for investigating
93 the effects of 17 β -trenbolone on reproductive behaviour. The guppy is a small, live-
94 bearing freshwater fish native to north-eastern South America that has a widespread
95 global distribution, precipitated by numerous deliberate and accidental introductions
96 (Lindholm et al., 2005). Importantly, throughout their range, guppies are known to
97 inhabit water bodies receiving agricultural waste (e.g., López-Rojas and Bonilla-
98 Rivero, 2000; Widianarko et al., 2000). Male guppies employ two alternate mating
99 strategies, either soliciting copulations from females through courtship (‘sigmoid
100 displays’) or coercing copulations through unsolicited ‘sneaking’ behaviour (Luyten

101 and Liley, 1991). The latter involves males surreptitiously approaching females from
102 behind to insert their gonopodium (a modified anal fin serving as an intromittent
103 organ) into the female's genital pore (Luyten and Liley, 1991). Female guppies are
104 choosy and are known, for example, to prefer males possessing greater orange
105 pigmentation (Houde, 1987). By preferentially associating with certain males over
106 others, females are able to directly influence mating outcomes (Shohet and Watt,
107 2004).

108 In this study we investigate the impacts of short-term (21-day) exposure to an
109 environmentally relevant concentration of 17β -trenbolone (22 ng/L) on male and
110 female reproductive behaviour in guppies. A short-term exposure duration was
111 employed as agricultural pollutants often enter aquatic habitats in pulses and these
112 temporally discrete contamination events can have persistent consequences (García et
113 al., 2011; Morthorst et al., 2010).

114

115 **Materials and methods**

116 *Ethical statement*

117 The research detailed in this paper was approved by the Biological Sciences Animal
118 Ethics Committee of Monash University (permit number: BSCI/2013/09) and
119 complies with all relevant State and Federal laws of Australia.

120

121 *Animal housing*

122 This study used laboratory-reared descendants of wild caught guppies from Alligator
123 Creek (19°26'17.94" S, 146°57'1.09" E), Queensland, Australia. Guppies were
124 separated by sex into 81 L housing tanks (60 cm × 45 cm × 30 cm) and acclimated to

125 laboratory conditions (25–27°C; 12:12 h light regime) for 2 months. Fish were fed
126 once daily (Otohime Hirame larval diet; 580–910 µm).

127

128 *Exposure set-up and monitoring*

129 A flow-through exposure design was used, as described by Saaristo et al. (2013). Fish
130 were assigned to identical 54 L separate-sex aquaria (60 cm × 30 cm × 30 cm), which
131 were monitored for temperature (\bar{x} = 26.38°C, SD = 0.52°C) and flow-through rates
132 (\bar{x} = 18.88 ml/min, SD = 0.59 ml/min). In total, 308 fish were randomly assigned to
133 one of seven 17β-trenbolone-exposure tanks, or one of seven unexposed tanks
134 containing fresh water (22 fish per tank).

135 The concentration of 17β-trenbolone used (\bar{x} = 22 ng/L, SD = 14.55 ng/L, n =
136 28) was monitored following Saaristo et al. (2013), with some modifications, using a
137 commercial Enzyme-Linked Immunosorbent Assay (ELISA). Weekly water samples
138 were drawn according to the protocol detailed by Saaristo et al. (2013).

139

140 *Behaviour trials*

141 To investigate the impact of 17β-trenbolone on the reproductive behaviour of guppies,
142 four treatments were employed: (1) unexposed male paired with unexposed female
143 (control; hereafter UU; n = 18), (2) unexposed male with exposed female (UE; n =
144 19), (3) exposed male with unexposed female (EU; n = 18), and (4) exposed male
145 with exposed female (EE; n = 20). Behavioural trials (n = 75) took place in 54 L
146 observation tanks (60 cm × 30 cm × 30 cm) containing fresh water. Trials involved a
147 five-minute period of acclimation, before both fish were released from holding
148 containers and allowed to freely interact, while their behaviour was video-recorded
149 for 15 minutes. Fish were euthanized immediately after trials using an overdose (40

150 mg/L) of anaesthetic clove oil, before morphological and colouration analyses were
151 conducted.

152 Reproductive behaviours were quantified from video recordings using the
153 event-recording software JWatcher V1.0 (Blumstein and Daniel, 2007). For males, we
154 quantified the number and total duration of courtship bouts performed (i.e., male
155 orienting in front of the female and performing courtship displays), as well as the
156 number of sneaking attempts (i.e., male surreptitiously approaching the female from
157 behind for forced copulation). Subsequently, we quantified the total time spent by
158 females actively associating with the male, a frequently used measure of mating intent
159 in poeciliid fishes (e.g., Kahn et al., 2010), including guppies (e.g., Shohet and Watt,
160 2004).

161

162 *Morphological analysis*

163 After behavioural trials, we measured the length of males and females (± 0.01 mm).
164 Males were also weighed (± 0.0001 g), and an index of male condition was derived
165 from a regression of the mass (g) of all males against their standard length (mm). This
166 male Condition Index was calculated as the residuals from the least squares regression
167 line (i.e., $\text{weight} = -0.164 + 0.016 \times \text{length}$).

168

169 *Colouration analysis*

170 Because female guppies typically prefer males with greater orange pigmentation (e.g.,
171 Houde, 1987), including in the source population of fish used in the present study
172 (Brooks and Endler, 2001; Gamble et al., 2003), the percentage of males' body
173 surface containing orange pigments was assessed using photographic colouration
174 analysis, performed immediately after behavioural trials. After euthanasia, fish were

175 photographed on the right side in a standardised fashion (Nikon D90, shutter speed =
176 1/250, Nikon AF Micro-Nikkor 60 mm f/2.8D).

177 Colouration analysis involved using Photoshop (CS5 Version 12.0 Extended)
178 to isolate the fishes' body surface, from snout to caudal peduncle (i.e., excluding fins).
179 Eight reference specimens were randomly selected (4 exposed, 4 unexposed).
180 Photoshop's Colour Range tool was used to sample the orange pigmentation of the
181 reference fish to create an orange pigmentation colour standard, which was applied to
182 all photographs. For each fish, the extent of orange pigmentation was calculated as the
183 number of orange pixels (i.e., pixels with colours belonging to the orange
184 pigmentation colour standard) as a proportion of the total body area (i.e., the number
185 of pixels forming the body surface).

186

187 *Statistical analysis*

188 Data were analysed in R version 3.0.2 (R Core Team, 2013). Where appropriate, data
189 were tested for normality (Shapiro-Wilk test, *shapiro.test* function; Royston, 1995)
190 and homogeneity of variance (Fligner-Killeen test, *fligner.test* function; Conover et
191 al., 1981). To assess whether exposure to 17 β -trenbolone impacted male Condition
192 Index, a two-sample *t*-test was used. Relationships were examined between
193 behavioural responses and a small suite of predictors through Multiple Linear
194 Regression (MLR) and by fitting Poisson Generalized Linear Models (GLMs)
195 (detailed in Supplementary materials). Vuong tests (*vuong* function, *pscl* package;
196 Jackman, 2012; Vuong, 1989) indicated zero-inflation of count data used for the
197 latter, which was accommodated by fitting Poisson models with the *zeroinfl* function
198 (*pscl* package; Zeileis et al., 2008). Continuous predictors were centred and
199 standardised to have zero mean and unit variance to enable direct comparison of their

200 coefficients. First-order interactions between predictors were pruned iteratively
201 through reverse stepwise elimination, leaving only those that were statistically
202 significant. *Post-hoc* evaluation of the differences in mean response across factor
203 levels (holding other predictors at their means) was performed through General Linear
204 Hypothesis Testing (GLHT) (*glht* function, *multcomp* package; Hothorn et al., 2008).
205 This process used partial Wald tests to assess differences between factor levels for
206 zero-inflated Poisson GLMs, and two-tailed *t*-tests to achieve the same for MLR
207 models.

208

209 **Results**

210 *Male behaviour*

211 Total male courting time varied significantly with both treatment and male Condition
212 Index. Relative to males in the control treatment (UU), both unexposed and exposed
213 males spent less time courting exposed females (*t*-test: $t = -2.177$, $df = 65$, $p = 0.033$
214 and $t = -2.190$, $df = 65$, $p = 0.032$, respectively; Fig. 1A). The effect of male
215 Condition Index on time spent courting depended on treatment, with a one standard
216 deviation increase in Condition Index (i.e., 0.012) relating to an increase in time spent
217 courting of 80.0 ± 28.8 seconds (*t*-test: $t = 2.782$, $df = 65$, $p = 0.007$) for males in the
218 control treatment. No relationship was detected between Condition Index and time
219 spent courting in any other treatment (*t*-test: all $p > 0.05$).

220 The number of courting events performed by males was associated with
221 treatment, male orange pigmentation, male Condition Index and female length.
222 Relative to males in the control treatment, both unexposed and exposed males
223 performed fewer courting events when paired with exposed females (partial Wald
224 test: $z = -2.219$, $p = 0.027$ and $z = -2.409$, $p = 0.016$, respectively; Fig. 1B). Increased

225 male orange pigmentation was associated with an increase in courting event
226 occurrence (partial Wald test: $z = 3.633$, $p < 0.001$; Fig. 2). Male Condition Index was
227 also positively associated with number of male courting events (partial Wald test: $z =$
228 7.509 , $p < 0.001$), with a one standard deviation increase in male Condition Index
229 (i.e., 0.012) yielding a 0.672 ± 0.435 increase in the log of the number of courting
230 events (on average, an additional 95.82% courting events per trial).

231 The number of sneaking attempts performed by males varied significantly
232 with treatment, male Condition Index and female length. The number of sneaking
233 attempts performed was significantly different between all pairs of treatments (Fig.
234 1C). Males in the control treatment performed fewer sneaking attempts than males in
235 any other treatment group (partial Wald test: all $z \geq 3.932$, all $p < 0.001$; Fig. 1C).
236 Unexposed males paired with exposed females snuck more than did exposed males
237 paired with exposed females (partial Wald test: $z = 2.704$, $p = 0.007$), with males from
238 both of these treatments performing more sneaking attempts than exposed males
239 paired with unexposed females (partial Wald test: $z = 5.127$, $p < 0.001$ and $z = 2.777$,
240 $p = 0.005$, respectively; Fig. 1C). Male Condition Index was negatively associated
241 with the number of sneaking attempts performed by males (partial Wald test: $z =$
242 -5.261 , $p < 0.001$), with a one standard deviation increase in Condition Index (i.e.,
243 0.012) associated with 0.223 ± 0.050 fewer sneaking attempts per trial (on the log
244 scale; corresponding to a 19.99% decrease in sneaking attempts). Female total length
245 also related negatively with the number of male sneaking attempts performed (partial
246 Wald test: $z = -6.159$, $p < 0.001$), with a one standard deviation increase in female
247 length (5.55 mm) yielding 0.340 ± 0.055 fewer sneaking attempts (on the log scale;
248 equivalent to 28.85% fewer sneaking attempts).

249

250 *Female behaviour*

251 Female association time was not affected by treatment (*t*-test: all $p > 0.05$). A reduced
252 model, excluding treatment, revealed significant effects of male orange pigmentation
253 (*t*-test: $t = 2.912$, $df = 71$, $p = 0.005$) and female length (*t*-test: $t = -6.298$, $df = 71$, $p <$
254 0.001) on female association time, but no effect of male Condition Index (*t*-test: $t =$
255 -0.186 , $df = 71$, $p = 0.853$). An increase in male orange pigmentation of one standard
256 deviation (i.e., 4.73%) corresponded with 33.3 ± 11.4 additional seconds spent by
257 females associating (Fig. 3). Longer females also spent less time associating with
258 males, with a one standard deviation increase in total female length (i.e., 5.6 mm)
259 relating with a decrease in female association time of 73.8 ± 11.7 seconds.

260

261 *Morphology*

262 Exposed males had, on average, a significantly higher Condition Index than
263 unexposed males (two-sample *t*-test: $t = -2.454$, $df = 70.174$, $p = 0.017$; Fig. 4). This
264 was due to exposed males being heavier (two-sample *t*-test: $t = -2.296$, $df = 72.985$, p
265 $= 0.025$), while male length was unaffected by exposure (two-sample *t*-test: $t =$
266 -1.103 , $df = 69.617$, $p = 0.274$).

267

268 **Discussion**

269 This research is the first to document altered male reproductive behaviour following
270 exposure to 17β -trenbolone at an environmentally relevant concentration. Males
271 paired with exposed females spent less time courting, and performed fewer courtship
272 bouts, than did males in the control treatment. Further, male exposure to 17β -
273 trenbolone led to an increase in sneaking behaviour when paired with unexposed
274 females. However, this finding was reversed when males were paired with exposed

275 females, with unexposed males sneaking more than exposed males. In addition,
276 regardless of male exposure status, males performed more sneaking behaviour when
277 paired with exposed females. More generally, males possessing greater areas of
278 orange pigmentation performed more courting bouts toward females than less-
279 colourful males, regardless of contamination with 17 β -trenbolone. This correlation
280 was anticipated, as orange pigmentation and display rate are both honest signals of
281 male condition (Kodric-Brown and Nicoletto, 2001; Nicoletto, 1993).

282 Vertebrate male sexual behaviours are reliant on androgens for their
283 production and maintenance (Cunningham et al., 2012). As a potent androgen agonist,
284 17 β -trenbolone has the capacity to disturb gonadal steroid signalling pathways by
285 increasing androgen synthesis. Given that a central role of androgens is the
286 modulation of male sexual and aggressive behaviours (Cunningham et al., 2012), the
287 anomalous presence of androgenic EDCs may result in the ‘hyper-masculinisation’ of
288 these traits in males. This phenomenon has been documented previously. For
289 example, exposure of African clawed frogs (*Xenopus laevis*) to androgenic endocrine
290 disruptors intensified androgen-dependent male mate calling (Hoffmann and Kloas,
291 2012), and increased the intensity of male sexual behaviours in various cyprinid fish
292 species (Belanger et al., 2010). The present results demonstrate that, in guppies, the
293 relative use of alternate reproductive strategies by males can be altered by exposure to
294 17 β -trenbolone. Specifically, despite 17 β -trenbolone having no significant effect on
295 mate solicitation by males (i.e., courtship), the increased number of sneaking attempts
296 performed by exposed males toward unexposed females (relative to the control group)
297 suggests that exposed males may favour this coercive reproductive strategy.
298 Interestingly, this finding was reversed given female exposure, with unexposed males
299 sneaking upon exposed females more than exposed males, possibly indicating a

300 greater capacity of unexposed males to take advantage of female exposure (although
301 the mechanisms underlying this possible phenomenon are not presently considered).

302 Disruption of the relative usage of alternative male reproductive strategies has
303 implications for male reproductive success, as sneaking behaviour is associated with
304 reduced insemination efficiency relative to copulations preceded by courtship
305 (Pilastro and Bisazza, 1999). Although sneaking behaviour is a viable sperm transfer
306 method, sperm transfer rates are approximately three times higher when delivered
307 after courtship (Pilastro and Bisazza, 1999). Further, postcopulatory female choice
308 may hamper the average reproductive success of males engaging in sneaking
309 behaviour. Such directional postcopulatory sexual selection has been documented in
310 female guppies, which have been shown to bias fertilisation in favour of more
311 colourful males (Pilastro et al., 2004).

312 Female exposure to 17β -trenbolone led to a decrease in the total duration and
313 frequency of male courtship behaviour, and an increase in male sneaking behaviour
314 (relative to the control group). Male guppies typically have very high levels of sexual
315 activity, meaning that females receive continual mating attempts that are mostly
316 unwanted (Houde, 1997). Typically, females actively avoid these incessant mating
317 attempts by swimming away from pursuing males (Houde, 1997). The present
318 findings suggest that the impact of female exposure to 17β -trenbolone on male
319 reproductive behaviour was male-driven, as female association time (indicative of
320 receptivity) was not influenced by exposure. This implies that exposure to 17β -
321 trenbolone inhibited the ability of females to actively avoid males. Regarding
322 courtship-initiated mating, females are able to exercise mating preferences — via both
323 precopulatory and postcopulatory mechanisms — by choosing which males to mate
324 with, and which sperm to use for fertilisation (Pilastro et al., 2004; Shohet and Watt,

325 2004). Sneaking behaviour, however, circumvents female precopulatory mate choice,
326 thus 17 β -trenbolone-exposure may directly interfere with sexual selection in guppies.

327 No significant effect of 17 β -trenbolone exposure on female reproductive
328 behaviour was detected. This result was unexpected as androgen receptors are not
329 sex-specific and endogenous androgens, as well as having an essential role in the
330 development and maintenance of male traits, serve important functions in female
331 vertebrates (Staub and De Beer, 1997). Androgens are involved in the regulation of
332 female sexual and aggressive behaviours (Staub and De Beer, 1997), meaning that the
333 mechanisms controlling these behaviours are particularly vulnerable to endocrine
334 disruption. Female exposure to androgenic EDCs has previously been linked with
335 physical and behavioural masculinisation, including the production of male
336 reproductive behaviour (Howell et al., 1980). Further, exposure of fathead minnow
337 (*Pimephales promelas*) to 17 β -trenbolone has been shown to severely alter female
338 reproductive biology and suppress the production of endogenous sex steroid hormones,
339 indicating masculinisation (Ankley et al., 2003). As such, 17 β -trenbolone was
340 expected to reduce female receptivity in the present study, but this was not observed.
341 The resilience of the metrics of female reproductive behaviour presently considered to
342 the concentration of 17 β -trenbolone employed suggests a differential vulnerability to
343 this EDC between sexes, a phenomenon previously documented in response to other
344 EDCs (e.g., Kundakovic et al., 2012). Consistent with prior research, females
345 exhibited a strong preference for males possessing greater orange pigmentation (e.g.,
346 Pilastro et al., 2004), regardless of exposure to 17 β -trenbolone.

347 This study found that exposure to 17 β -trenbolone was associated with an
348 increase in male Condition Index, due to exposed males being heavier, despite there
349 being no significant difference in length between exposed and unexposed males. This

350 weight gain was anticipated, given the potent growth-promoting activity of 17 β -
351 trenbolone. A similar finding was reported in a study exposing juvenile guppies to
352 trenbolone acetate for 60 days. In that study, however, fish were exposed dietarily to
353 300 mg/kg of trenbolone acetate (Zamora et al., 2008), which is neither an
354 environmentally relevant mode, nor concentration, of exposure. The present research,
355 however, indicates that environmentally realistic levels of contamination are sufficient
356 to cause weight gain, even with short-term exposure. For various taxa, including fish,
357 heavier males have greater reproductive success in competitive breeding scenarios
358 (e.g., Jacob et al., 2009). Thus, although the present study did not test for the effect of
359 17 β -trenbolone on male competitive ability, the increased weights of males exposed
360 to 17 β -trenbolone may confer an advantage in jockeying for contested fertilisations.
361 This potential scenario holds ecological relevance, as EDC concentrations are
362 typically spatially and temporally variable (e.g., Grover et al., 2011), meaning that
363 interactions between exposed and unexposed individuals are likely.

364

365 **Conclusion**

366 This study reports that short-term (21-day) exposure to an environmentally relevant
367 concentration (22 ng/L) of the androgenic endocrine disruptor 17 β -trenbolone can
368 alter reproductive behaviour and morphology in the guppy. This is the first study to
369 show altered reproductive behaviours in male animals resulting from an
370 environmentally realistic exposure to 17 β -trenbolone. Given the prevalence and
371 potent biological activity of 17 β -trenbolone, the ongoing multidisciplinary scrutiny of
372 this EDC is necessary to reveal the consequences of its presence in the environment.

373

374

375 **Acknowledgements**

376 We thank John Endler and his research group for supplying fish for this study.

377

378 **Funding statement**

379 This study was supported by a Discovery Grant from the Australian Research Council
380 (DP130100385) (awarded to BMW) and an Academy of Finland Postdoctoral
381 Researcher Fellowship (265629) (to MS). At the initiation of this study, the Centre for
382 Aquatic Pollution Identification and Management (CAPIM) received foundation
383 funding from the Victorian Science Agenda Investment Fund managed by the
384 Department of Business and Innovation (DBI) (www.innovation.vic.gov.au), with
385 additional funding from Melbourne Water and the Department of Environment and
386 Primary Industries (Victoria).

387

388 **Appendix A. Supplementary materials**

389 Supplementary data to this article can be found online at

390

391

392

393

394

395

396

397

398

399

400 **References**

- 401 Ankley, G.T., Jensen, K.M., Makynen, E.A., Kahl, M.D., Korte, J.J., Hornung, M.W.,
402 Henry, T.R., Denny, J.S., Leino, R.L., Wilson, V.S., Cardon, M.C., Hartig,
403 P.C., Gray, L.E., 2003. Effects of the androgenic growth promoter 17- β -
404 trenbolone on fecundity and reproductive endocrinology of the fathead
405 minnow. *Environ. Toxicol. Chem.* 22, 1350–1360.
406 [http://dx.doi.org/10.1897/1551-5028\(2003\)022<1350:EOTAGP>2.0.CO;2](http://dx.doi.org/10.1897/1551-5028(2003)022<1350:EOTAGP>2.0.CO;2)
- 407 Belanger, R.M., Pachkowski, M.D., Stacey, N.E., 2010. Methyltestosterone-induced
408 changes in electro-olfactogram responses and courtship behaviors of
409 cyprinids. *Chem. Senses* 35, 65–74. <http://dx.doi.org/10.1093/chemse/bjp085>
- 410 Blumstein, D.T., Daniel, J.C., 2007. *Quantifying Behaviour the JWatcher Way*.
411 Sinauer Associates Inc., Sunderland.
- 412 Brooks, R., Endler, J.A., 2001. Direct and indirect sexual selection and quantitative
413 genetics of male traits in guppies (*Poecilia reticulata*). *Evolution* 55, 1002–
414 1015. [http://dx.doi.org/10.1554/0014-3820\(2001\)055\[1002:DAISSA\]2.0.CO;2](http://dx.doi.org/10.1554/0014-3820(2001)055[1002:DAISSA]2.0.CO;2)
- 415 Candolin, U., Wong, B.B.M., 2012. *Behavioural Responses to a Changing World:
416 Mechanisms and Consequences*. Oxford University Press, Oxford.
- 417 Colborn, T., vom Saal, F.S., Soto, A.M., 1993. Developmental effects of endocrine-
418 disrupting chemicals in wildlife and humans. *Environ. Health Perspect.* 101,
419 378–384. <http://dx.doi.org/10.2307/3431890>
- 420 Conover, W.J., Johnson, M.E., Johnson, M.M., 1981. A comparative study of tests for
421 homogeneity of variances, with applications to the outer continental shelf
422 bidding data. *Technometrics* 23, 351–361. <http://dx.doi.org/10.2307/1268225>

423 Core Team, R., 2013. R: A Language and Environment for Statistical Computing. R
424 Foundation for Statistical Computing, Vienna, Austria ([http://www.R-](http://www.R-project.org/)
425 [project.org/](http://www.R-project.org/)).

426 Cunningham, R.L., Lumia, A.R., McGinnis, M.Y., 2012. Androgen receptors, sex
427 behavior, and aggression. *Neuroendocrinology* 96, 131–140.
428 <http://dx.doi.org/10.1159/000337663>

429 Diamanti-Kandarakis, E., Bourguignon, J.P., Giudice, L.C., Hauser, R., Prins, G.S.,
430 Soto, A.M., Zoeller, R.T., Gore, A.C., 2009. Endocrine-disrupting chemicals:
431 an Endocrine Society scientific statement. *Endocr. Rev.* 30, 293–342.
432 <http://dx.doi.org/10.1210/er.2009-0002>

433 Durhan, E.J., Lambright, C.S., Makynen, E.A., Lazorchak, J., Hartig, P.C., Wilson,
434 V.S., Gray, L.E., Ankley, G.T., 2006. Identification of metabolites of
435 trenbolone acetate in androgenic runoff from a beef feedlot. *Environ. Health*
436 *Perspect.* 114, 65–68. <http://dx.doi.org/10.1289/ehp.8055>

437 Gall, H.E., Sassman, S.A., Lee, L.S., Jafvert, C.T., 2011. Hormone discharges from a
438 midwest tile-drained agroecosystem receiving animal wastes. *Environ. Sci.*
439 *Technol.* 45, 8755–8764. <http://dx.doi.org/10.1021/es2011435>

440 Gamble, S., Lindholm, A.K., Endler, J.A., Brooks, R., 2003. Environmental variation
441 and the maintenance of polymorphism: the effect of ambient light spectrum on
442 mating behaviour and sexual selection in guppies. *Ecol. Lett.* 6, 463–472.
443 <http://dx.doi.org/10.1046/j.1461-0248.2003.00449.x>

444 García, J.H., Heberling, M.T., Thurston, H.W., 2011. Optimal pollution trading
445 without pollution reductions: a note. *J. Am. Water Resour. Assoc.* 47, 52–58.
446 <http://dx.doi.org/10.1111/j.1752-1688.2010.00476.x>

447 Grover, D.P., Balaam, J., Pacitto, S., Readman, J.W., White, S., Zhou, J.L., 2011.
448 Endocrine disrupting activities in sewage effluent and river water determined
449 by chemical analysis and in vitro assay in the context of granular activated
450 carbon upgrade. *Chemosphere* 84, 1512–1520.
451 <http://dx.doi.org/10.1016/j.chemosphere.2011.04.032>

452 Hoffmann, F., Kloas, W., 2012. Effects of environmentally relevant concentrations of
453 the xeno-androgen, methyl dihydrotestosterone, on male and female mating
454 behavior in *Xenopus laevis*. *Chemosphere* 87, 1246–1253.
455 <http://dx.doi.org/10.1016/j.chemosphere.2012.01.030>

456 Hothorn, T., Bretz, F., Westfall, P., 2008. Simultaneous inference in General
457 Parametric Models. *Biom. J.* 50, 346–363.
458 <http://dx.doi.org/10.1002/bimj.200810425>

459 Houde, A.E., 1987. Mate choice based upon naturally occurring color-pattern
460 variation in a guppy population. *Evolution* 41, 1–10.
461 <http://dx.doi.org/10.2307/2408968>

462 Houde, A.E., 1997. Sex, Color, and Mate Choice in Guppies. Princeton University
463 Press, Princeton.

464 Howell, W.M., Black, D.A., Bortone, S.A., 1980. Abnormal expression of secondary
465 sex characters in a population of mosquitofish, *Gambusia affinis holbrooki*:
466 evidence for environmentally-induced masculinization. *Copeia* 4, 676–681.
467 <http://dx.doi.org/10.2307/1444443>

468 Jackman, S., 2012. pscl: Classes and Methods for R Developed in the Political
469 Science Computational Laboratory, Stanford University. Department of
470 Political Science, Stanford University, Stanford.

471 Jacob, A., Evanno, G., Renai, E., Sermier, R., Wedekind, C., 2009. Male body size
472 and breeding tubercles are both linked to intrasexual dominance and
473 reproductive success in the minnow. *Anim. Behav.* 77, 823–829.
474 <http://dx.doi.org/10.1016/j.anbehav.2008.12.006>

475 Kahn, A.T., Mautz, B., Jennions, M.D., 2010. Females prefer to associate with males
476 with longer intromittent organs in mosquitofish. *Biol. Lett.* 6, 55–58.
477 <http://dx.doi.org/10.1098/rsbl.2009.0637>

478 Kodric-Brown, A., Nicoletto, P.F., 2001. Female choice in the guppy (*Poecilia*
479 *reticulata*): the interaction between male color and display. *Behav. Ecol.*
480 *Sociobiol.* 50, 346–351. <http://dx.doi.org/10.1007/s002650100374>

481 Kolodziej, E.P., Qu, S., Forsgren, K.L., Long, S.A., Gloer, J.B., Jones, G.D., Schlenk,
482 D., Baltrusaitis, J., Cwiertny, D.M., 2013. Identification and environmental
483 implications of photo-transformation products of trenbolone acetate
484 metabolites. *Environ. Sci. Technol.* 47, 5031–5041.
485 <http://dx.doi.org/10.1021/es3052069>

486 Kundakovic, M., Gudsnuk, K., Franks, B., Madrid, J., Miller, R.L., Perera, F.P.,
487 Champagne, F.A., 2012. Sex-specific epigenetic disruption and behavioral
488 changes following low-dose in utero bisphenol A exposure. *Proc. Natl. Acad.*
489 *Sci. U. S. A.* 110, 9956–9961. <http://dx.doi.org/10.1073/pnas.1214056110>

490 Lindholm, A.K., Breden, F., Alexander, H.J., Chan, W.K., Thakurta, S.G., Brooks, R.,
491 2005. Invasion success and genetic diversity of introduced populations of
492 guppies *Poecilia reticulata* in Australia. *Mol. Ecol.* 14, 3671–3682.
493 <http://dx.doi.org/10.1111/j.1365-294X.2005.02697.x>

494 López-Rojas, H., Bonilla-Rivero, A.L., 2000. Anthropogenically induced fish
495 diversity reduction in Lake Valencia Basin, Venezuela. *Biodivers. Conserv.* 9,
496 757–765. <http://dx.doi.org/10.1023/A:1008945813101>

497 Luyten, P.H., Liley, N.R., 1991. Sexual selection and competitive mating success of
498 male guppies (*Poecilia reticulata*) from four Trinidad populations. *Behav.*
499 *Ecol. Sociobiol.* 28, 329–336. <http://dx.doi.org/10.1007/BF00164382>

500 Melvin, S.D., Wilson, S.P., 2013. The utility of behavioral studies for aquatic
501 toxicology testing: a meta-analysis. *Chemosphere* 93, 2217–2223.
502 <http://dx.doi.org/10.1016/j.chemosphere.2013.07.036>

503 Morthorst, J.E., Holbech, H., Bjerregaard, P., 2010. Trenbolone causes irreversible
504 masculinization of zebrafish at environmentally relevant concentrations.
505 *Aquat. Toxicol.* 98, 336–343. <http://dx.doi.org/10.1016/j.aquatox.2010.03.008>

506 Neumann, F., 1976. Pharmacological and endocrinological studies on anabolic agents.
507 *Environ. Qual. Saf. Suppl.* 5, 253–264.

508 Nicoletto, P.F., 1993. Female sexual response to condition-dependent ornaments in
509 the guppy, *Poecilia reticulata*. *Anim. Behav.* 46, 441–450.
510 <http://dx.doi.org/10.1006/anbe.1993.1213>

511 Pilastro, A., Bisazza, A., 1999. Insemination efficiency of two alternative male mating
512 tactics in the guppy (*Poecilia reticulata*). *Proc. R. Soc. B* 266, 1887–1891.
513 <http://dx.doi.org/10.1098/rspb.1999.0862>

514 Pilastro, A., Simonato, M., Bisazza, A., Evans, J.P., 2004. Cryptic female preference
515 for colorful males in guppies. *Evolution* 58, 665–669.
516 <http://dx.doi.org/10.1111/j.0014-3820.2004.tb01690.x>

517 Royston, P., 1995. Remark AS R94: a remark on algorithm AS 181: the *W*-test for
518 normality. *J. R. Stat. Soc. Ser. C Appl. Stat.* 44, 547–551.
519 <http://dx.doi.org/10.2307/2986146>

520 Saaristo, M., Craft, J.A., Lehtonen, K.K., Björk, H., Lindström, K., 2009. Disruption
521 of sexual selection in sand gobies (*Pomatoschistus minutus*) by 17 α -ethinyl
522 estradiol, an endocrine disruptor. *Horm. Behav.* 55, 530–537.
523 <http://dx.doi.org/10.1016/j.yhbeh.2009.01.006>

524 Saaristo, M., Tomkins, P., Allinson, M., Allinson, G., Wong, B.B.M., 2013. An
525 androgenic agricultural contaminant impairs female reproductive behaviour in
526 a freshwater fish. *PLOS ONE* 8, e62782.
527 <http://dx.doi.org/10.1371/journal.pone.0062782>

528 Schiffer, B., Daxenberger, A., Meyer, K., Meyer, H.H.D., 2001. The fate of
529 trenbolone acetate and melengestrol acetate after application as growth
530 promoters in cattle: environmental studies. *Environ. Health Perspect.* 109,
531 1145–1151. <http://dx.doi.org/10.2307/3454862>

532 Schwarzenbach, R.P., Escher, B.I., Fenner, K., Hofstetter, T.B., Johnson, C.A., von
533 Gunten, U., Wehrli, B., 2006. The challenge of micropollutants in aquatic
534 systems. *Science* 313, 1072–1077. <http://dx.doi.org/10.1126/science.1127291>

535 Shohet, A.J., Watt, P.J., 2004. Female association preferences based on olfactory cues
536 in the guppy, *Poecilia reticulata*. *Behav. Ecol. Sociobiol.* 55, 363–369.
537 <http://dx.doi.org/10.1007/s00265-003-0722-0>

538 Staub, N.L., De Beer, M., 1997. The role of androgens in female vertebrates. *Gen.*
539 *Comp. Endocrinol.* 108, 1–24. <http://dx.doi.org/10.1006/gcen.1997.6962>

540 Vuong, Q.H., 1989. Likelihood ratio tests for model selection and non-nested
541 hypotheses. *Econometrica* 57, 307–333. <http://dx.doi.org/10.2307/1912557>

542 Widianarko, B., Van Gestel, C.A.M., Verweij, R.A., Van Straalen, N.M., 2000.
543 Associations between trace metals in sediment, water, and guppy, *Poecilia*
544 *reticulata* (Peters), from urban streams of Semarang, Indonesia. *Ecotoxicol.*
545 *Environ. Saf.* 46, 101–107. <http://dx.doi.org/10.1006/eesa.1999.1879>
546 Zamora, H.S., Hernández, A.A., Herrera, S.M., Peña, E.M., 2008. Anabolic and
547 androgenic effect of steroid trenbolone acetate on guppy (*Poecilia reticulata*).
548 *Vet. México* 39, 269–277.
549 Zeileis, A., Kleiber, C., Jackman, S., 2008. Regression models for count data in R. J.
550 *Stat. Softw.* 27, 1–25.
551
552
553
554
555
556
557
558
559
560
561
562
563
564
565
566
567

568 **Figure legends**

569 **Fig. 1.** Mean (\pm SE) of the (A) duration (seconds) of male courting behaviour
570 (orienting and sigmoid behaviours), (B) number of male courting events (orienting
571 and sigmoid behaviours) and (C) number of male sneaking attempts across treatment
572 groups ($n = 18, 19, 18$ and 20 , for UU, UE, EU and EE, respectively). Treatments
573 indicate unexposed (U) and exposed (E) fish, with male treatment followed by female
574 treatment. Treatments without lower case letters in common are significantly
575 different.

576

577 **Fig. 2.** Expected number of male courting events given male orange pigmentation (%
578 of body area) and treatment ($n = 18, 19, 18$ and 20 , for UU, UE, EU and EE,
579 respectively). Treatments indicate unexposed (U) and exposed (E) fish, with male
580 treatment followed by female treatment.

581

582 **Fig. 3.** Female association time (seconds) given male orange pigmentation (% of body
583 area) across treatments ($n = 75$). Solid black line indicates least squares regression
584 line and shaded region indicates 95% confidence interval.

585

586 **Fig. 4.** Boxplots of male Condition Index for unexposed males ($n = 37$) and those
587 exposed to 17β -trenbolone ($n = 38$).