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Adaptation to reef habitats through selection on the coral animal and its associated microbiome

Short title: Local adaptation in corals

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29 **Abstract**

30 Spatially adjacent habitats on coral reefs can represent highly distinct environments, often
31 harbouring different coral communities. Yet, certain coral species thrive across divergent
32 environments. It is unknown whether the forces of selection are sufficiently strong to overcome
33 the counteracting effects of the typically high gene flow over short distances, and for local
34 adaptation to occur. We screened the coral genome (using restriction-site-associated sequencing
35 [RAD-seq]), and characterized both the dinoflagellate photosymbiont and tissue-associated
36 prokaryote microbiomes (using metabarcoding) of a reef flat and slope population of the reef-
37 building coral, *Pocillopora damicornis*, at two locations on Heron Island in the southern Great
38 Barrier Reef. Reef flat and slope populations were separated by <100 m horizontally and ~5 m
39 vertically and the two study locations were separated by ~1 km. For the coral host, genetic
40 divergence between habitats was much greater than between locations, suggesting limited gene
41 flow between the flat and slope populations. Consistent with environmental selection, outlier loci
42 primarily belonged to the conserved, minimal cellular stress response, likely reflecting adaptation
43 to the different temperature and irradiance regimes on the reef flat and slope. Similarly, the
44 prokaryote community differed across both habitat and, to a lesser extent, location, whereas the
45 dinoflagellate photosymbionts differed by habitat but not location. The observed intra-specific
46 diversity associated with divergent habitats supports that environmental adaptation involves
47 multiple members of the coral holobiont. Adaptive alleles or microbial associations present in

48 coral populations from the environmentally-variable reef flat may provide a source of adaptive
49 variation for assisted evolution approaches, through assisted gene flow, artificial cross-breeding
50 or probiotic inoculations, with the aim to increase climate resilience in the slope populations.

51

52 **Introduction**

53 Adaptation to local environmental conditions is a common phenomenon in both the terrestrial
54 and the marine environment (Sanford & Kelly 2011; Shafer & Wolf 2013), including coral reef
55 ecosystems (Bay & Palumbi 2014; Dixon *et al.* 2015; Edmunds *et al.* 2012). Local adaptation in
56 reef-building corals has occurred over a range of spatial scales, i.e., between regions (Riegl *et al.*
57 2011), between adjacent seas (Howells *et al.* 2016), along latitudinal gradients (Berkelmans &
58 van Oppen 2006), across the continental shelf (Berkelmans & van Oppen 2006), and between
59 habitats on the same reef (Bay & Palumbi 2014). Compared to reef slope habitats, reef flat or
60 back reef habitats are far more variable in temperature, light and partial pressure of carbon
61 dioxide ($p\text{CO}_2$) (Albright *et al.* 2015; Ohde & van Woesik 1999; Price *et al.* 2012; Santos *et al.*
62 2011; Shaw *et al.* 2012). Corals from environmentally variable locations are generally better able
63 to withstand stress through phenotypic plasticity (Barshis *et al.* 2013; Kenkel & Matz 2016;
64 Oliver & Palumbi 2011) and have greater adaptive potential (McClanahan *et al.* 2007) compared
65 to those that live in more stable environments. The potential of such populations to serve as a
66 source for adaptive genetic variation for surrounding habitats remains poorly explored.
67 Potentially, these populations can provide adaptive alleles that could be harnessed in assisted
68 gene flow or selective breeding initiatives that aim to increase climate resilience in the receiving
69 populations (Hoegh-Guldberg *et al.* 2008; van Oppen *et al.* 2017; van Oppen *et al.* 2015).

70 Reef-building corals harbour a diverse microbiome comprised of bacteria, archaea, dinoflagellate
71 photosymbionts (*Symbiodinium* spp.), fungi, chlorophytes, viruses and other micro-organisms
72 (Blackall *et al.* 2015). It is well known that the endosymbiotic *Symbiodinium* community plays an
73 important role in coral nutrition and calcification, as well as the ability of corals to tolerate and
74 recover from environmental stress (Bay *et al.* 2016; Berkelmans & van Oppen 2006), with not all
75 *Symbiodinium* types providing the same benefits to the host (van Oppen *et al.* 2009). Prokaryotes

76 are critical for nitrogen fixation, sulphur cycling, and coral host immunity (Bourne *et al.* 2016;
77 Thompson *et al.* 2014). Differences in the prokaryotic community composition in conspecific
78 corals from distinct environments (Glasl *et al.* 2017a; Hernandez-Agreda *et al.* 2016a) may be a
79 signature of acclimatization and/or adaptation. Experimental heat stress induced changes in the
80 bacterial community composition of *Acropora hyacinthus* colonies acclimatized to a relatively
81 cooler and stable environment, but not in those that were acclimatized to a warmer and more
82 variable environment, suggesting coral-associated bacteria may contribute to coral thermal
83 tolerance in this species (Ziegler *et al.* 2017). Conversely, stable cnidarian-associated prokaryote
84 communities under simulated environmental change (Webster *et al.* 2016) or in different habitats
85 (Muller *et al.* 2016) have also been reported, and it is not yet clear to what extent the prokaryotic
86 community contributes to the acclimatization/adaptation of corals to different reef environments.

87 The brooding coral, *Pocillopora damicornis*, is a common and widespread reef species that
88 inhabits a wide range of reef habitats, including reef flats and deeper reef slopes. While
89 recruitment in this species is often localized, it is supplemented by dispersal of short- and long-
90 distance migrants (Torda *et al.* 2013a). It is not known whether the forces of selection on the reef
91 flat versus slope can overcome the homogenizing effects of gene flow and lead to local
92 adaptation between these spatially adjacent sites. Previous studies have demonstrated that
93 conspecific coral populations which inhabit different habitats can be genetically distinct (Benzie
94 *et al.* 1995; Bongaerts *et al.* 2015; Bongaerts *et al.* 2011; Bongaerts *et al.* 2010; Souter & Grahn
95 2008; Warner *et al.* 2015), but the underlying causes of these patterns have been difficult to
96 interpret (Torda *et al.* 2013b). Further, the genomic pathways and loci under selection were not
97 identified in these earlier studies (Bongaerts *et al.* 2013). Here, we examine the occurrence of
98 local adaptation in multiple compartments of the *P. damicornis* holobiont, between the
99 environmentally divergent, but spatially adjacent (<100 m) reef flat and reef slope habitats of *P.*
100 *damicornis* at Heron Island in the southern Great Barrier Reef (GBR). Specifically, divergence
101 was assessed at the level of (1) the coral host by using genome-wide single nucleotide
102 polymorphism (SNP) analysis, (2) the associated photosynthetic endosymbionts (*Symbiodinium*
103 spp.) by ribosomal DNA (rDNA) internal transcribed spacer 2 (ITS2) metabarcoding, and (3)
104 tissue-associated prokaryote members of the microbiome by 16S ribosomal RNA (rRNA)

105 metabarcoding. Our experimental design included two replicate locations, with 24 colonies
106 sampled from each of the reef flat and slope sites at the two locations that are ~1 km apart, Coral
107 Canyons and Coral Gardens.

108

109 **Materials and Methods**

110 *Sample collection, local seawater temperatures and light levels*

111 Fragments of 96 *P. damicornis* colonies (24 from each of four sites) were haphazardly collected
112 between 27/02/2012 and 11/03/2012 from the reef flat (1-2 m depth) and slope (4-6 m depth) at
113 two locations near the Heron Island Research Station on the Great Barrier Reef: “Coral Gardens”
114 (flat: 23°26’48.50”S, 151°54’46.34” E; slope: 23°26’50.54”S, 151°54’45.46” E) and “Coral
115 Canyons” (flat: 23°27’14.50”S, 151°55’34.55” E; slope: 23°27’18.80”, 151°55’28.29” E) under
116 Great Barrier Reef Marine Park Authority permit G12/34752.1. Samples were coded as follows:
117 GF = Coral Gardens flat [samples 101-124], GS = Coral Gardens slope [samples 201-224], CF =
118 Coral Canyons flat [samples 301-324], CS = Coral Canyons slope [samples 401-424]). Seawater
119 temperature was monitored at the collection sites (~30 cm above the substratum) every 10 min
120 between 03/03/2012 and 16/08/2013 using *in situ* temperature loggers (Sensus Ultra, ReefNet).
121 The temperature logger failed at the Coral Gardens sites in August 2012 and, hence, no data are
122 available from this site after this date. Photosynthetically active radiation (PAR) was measured
123 underwater at the collection sites (~30 cm above the substratum) between 9 am and 3 pm from
124 05/03/2012 – 02/04/2012 and 25/03/2013 - 16/08/2013 using integrating light sensors (Odyssey
125 Dataflow Systems). The light sensors were fitted with wipers designed to keep the instrument
126 clean from unwanted deposits such as mud and bio-fouling during deployment, a setup prone to
127 technical failure. Due to the short interval measured in 2012 these data were not included. Data
128 were only collected for four months at the Canyons slope site in 2013. Profiles were
129 indistinguishable between habitats when they could be compared and hence only the complete
130 2013 data series from the Garden location was displayed. Although not measured here, the
131 location of the Garden reef was more exposed and it is likely the corals here would experience
132 higher rates of water flow. Ocean chemistry was not measured, but pH and dissolved inorganic

133 carbon, dissolved oxygen and other parameters have been found to vary on daily cycles in the
134 Heron Island lagoon (Kline et al 2012; Santos et al 2011). We did not have access to equivalent
135 data sets for the slope and, hence, do not discuss this further except to note that the reef flat sites
136 were likely exposed to a range of variable environmental conditions.

137

138 *DNA extractions, RAD-seq library preparation and sequencing*

139 Total genomic DNA (gDNA) was extracted from ethanol-preserved specimens (n = 96) using a
140 salt-precipitation method including both RNase A and Proteinase K treatment (van Oppen *et al.*
141 2011). Total gDNA was normalised to around 75 ng.µL⁻¹ by fluorometry using the Quant-It
142 PicoGreen dsDNA quantification kit (Thermo Fisher), and by also assessing the high molecular
143 weight (HMW) band against an uncut lambda DNA standard on agarose gels. DNA purity was
144 determined spectrophotometrically from Nanodrop 1000 (Thermo Scientific) 260/230 and
145 260/280 nm absorbance ratios. RAD-seq library construction and sequencing was conducted at
146 Floragenex Inc. (Eugene, USA) following the protocol described by Baird *et al.* (2008). In short,
147 gDNA was digested using the SbfI restriction endonuclease (6 bp recognition sequence
148 TGCAGG), and “barcoded” by ligating sequencing adapters with a set of index nucleotides.
149 Reactions were pooled and sheared, after which they were size-selected and taken through
150 several purification steps. RAD-seq fragment libraries were sequenced using single read 100 bp
151 chemistry on the Illumina HiSeq 2000 platform using the manufacturer’s recommended protocol,
152 with a subset of the samples sequenced across multiple runs to increase coverage. RAD-seq data
153 are available as fastq files on NCBI Sequence Read Archive (accession number PRJNA450871),
154 and VCF files on Dryad (doi:10.5061/dryad.j09c05c).

155

156 *Coral ID assay*

157 All collected colonies were of the *P. damicornis* morphology (Schmidt-Roach *et al.* 2013), and
158 their species ID was confirmed by a molecular assay. The assay is based on PCR amplification of
159 a portion of the mitochondrial putative control region, which shows consistent and easily

160 identifiable fragment size differences after Alu1 restriction enzyme digestion of the amplicons
161 (Torda *et al.* 2013c).

162

163 *Identification of coral host loci*

164 Clustering of RAD-seq loci and SNP variant calling was carried out using PyRAD v3.0.6 (Eaton
165 2014) to allow for indel variation (common in marine invertebrates), using a minimum coverage
166 of 6, a clustering threshold of 88%, and a maximum of 4 low-quality sites in reads (using a
167 PHRED threshold of 20). Further parsing and analyses were carried out using custom Python
168 scripts (available through <https://github.com/pimbongaerts/radseq>) unless otherwise indicated.
169 One representative sequence was then extracted for each RAD-seq locus to identify only those
170 loci belonging to the coral host. Putative coral host loci that both mapped (using BWA; Li &
171 Durbin 2009) and matched (using BLASTn; max. e-value of 10^{-15}) against at least one of three *P.*
172 *damicornis* transcriptome references (Lundgren *et al.* 2013; Mayfield *et al.* 2014; Traylor-
173 Knowles *et al.* 2011) were retained. Remaining *Symbiodinium* contamination was identified
174 using a BLASTn comparison (max. e-value of 10^{-15}) against genomes of *Symbiodinium* B1
175 (Shoguchi *et al.* 2013) and *Symbiodinium* C1 (Liu *et al.* 2017) and transcriptomes of C and D
176 types (Ladner *et al.* 2012; Levin *et al.* 2016).

177

178 *QC of SNPs and identification of clones*

179 Given a slight overrepresentation of SNPs toward the end of reads, RAD-seq loci were truncated
180 to 80 bp. Remaining SNPs were evaluated for significant deviations from Hardy-Weinberg
181 equilibrium using Arlequin (arlecore v3.5.2.2; Excoffier & Lischer 2010), and loci with SNPs
182 that either deviated (excess or deficit) in at least three populations or had a significant excess in at
183 least one population (to filter out potential paralogous loci) were removed. Genetic similarity
184 (Hamming-based) was calculated between all individuals, and those with >94% similarity were
185 considered clone mates based on a distinct break in the frequency distribution of similarities
186 (using the "vcf_clone_detect" script). Two separate data sets were then compiled: one with clones
187 and one without clones. In both datasets, only SNPs with a minimum representation of 30% in

188 each population were retained in the data sets, and minor alleles that occurred as singleton or
189 doubleton were removed (to reduce elimination of non-erroneous minor alleles through a
190 frequency filter).

191

192 *Outlier analyses and putatively neutral SNPs*

193 Initially, outliers were identified for the overall dataset (considering only biallelic SNPs) and for
194 each pairwise population comparison using both Lositan (Antao *et al.* 2008) and BayeScan (Foll
195 2012), with a threshold of 0.99 and a false discovery rate (FDR) of 0.1. To conservatively
196 identify between-habitat outlier SNPs, i.e., SNPs that are under selection from different
197 environmental conditions in the two habitats, we selected those that met all of the following
198 criteria: (1) identified by both Bayescan and Fdist as “overall” outliers, (2) identified by Fdist as
199 outliers in at least two cross-habitat comparisons, and (3) never identified by Fdist as outliers in
200 within-habitat, between-location comparisons (i.e., between locations). This strategy provides a
201 high level of confidence in the identified loci being under selection due to habitat-specific
202 differences. Loci containing outliers were annotated by BLASTx search of the corresponding
203 transcriptome contig to ZoophyteBase (Dunlap *et al.* 2013). This database was used because it
204 has good annotations.

205 To identify a subset of putatively neutral SNPs, the outlier analyses were repeated using a cut-off
206 of 0.95 (lower) and a FDR of 0.1 (the same), and retaining SNPs that were never identified in any
207 of the overall analyses. Again, separate data sets were compiled for those with clones and without
208 clones.

209

210 *Genetic structure*

211 Genetic clustering and potential admixture was assessed using the non-clonal version of both the
212 overall and "neutral" datasets using the “structure_mp” script (Bongaerts *et al.* 2017) for
213 STRUCTURE (Pritchard *et al.* 2000). STRUCTURE was run for a K ranging from 2 to 6 using

214 the admixture model and correlated allele frequencies with no priors, an MCMC burn-in of
215 100,000 chains and 50,000 replications. Ten independent replicates were run for each K, using a
216 different subset of SNPs for each replicate (one per locus, sampled with replacement), with
217 results summarized in CLUMPP (Jakobsson & Rosenberg 2007), and estimators for the most
218 likely K calculated as in Puechmaille (2016) and Evanno et al. (2005). Pairwise genetic distances
219 (F_{ST}) were calculated in the “adegenet” R package (Jombart & Ahmed 2011).

220

221 *Prokaryote community composition*

222 DNA was diluted tenfold with MilliQ water prior to PCR. PCR was carried out in Applied
223 Biosystems 2720 thermal cyclers under the following cycling conditions: 1 cycle at 95°C for 15
224 min; 30 cycles of 95°C for 30 s, 55°C for 30 s, 72°C for 30 s; and one final extension cycle of
225 72°C for 7 min. Triplicate 15 µL reactions were set up for each sample. Each reaction consisted
226 of 7.5 µL Qiagen Multiplex PCR mastermix reagent, one µL of template DNA, and 4 pmol of
227 each of 16S rRNA gene V4 515F (5'-
228 TCGTCGGCAGCGTCAGATGTGTATAAGAGACAGGTGYCAGCMGCCGCGGTAA-3';
229 Apprill *et al.* 2015) and 806Rb (5'-
230 GTCTCGTGGGCTCGGAGATGTGTATAAGAGACAGGGACTACNVGGGTWTCTAAT-3';
231 Parada *et al.* 2015) primer. The underlined segments represent Illumina adapters (Illumina, San
232 Diego, CA, USA). Three µL of PCR product were run on a 2.5% agarose, 0.5x TBE gel to
233 confirm successful amplification. Ten µL from each of the triplicate PCRs for each sample were
234 pooled and sent to a sequencing service provider (Ramaciotti Centre for Genomics the University
235 of New South Wales) for PCR purification, amplicon library construction and Illumina MiSeq
236 analysis. Sequencing was conducted using Illumina MiSeq 300 bp paired-end chemistry.
237 Retrieved raw reads plus respective metadata have been deposited in the NCBI Sequence Read
238 Archive under BioProject accession number PRJNA380169. Demultiplexed reads were analysed
239 in QIIME (version 1.9.0) (Caporaso *et al.* 2010). Paired-end reads were merged into contigs,
240 quality trimmed using default thresholds (and a minimum Phred quality score of 25) and primers
241 removed. Chimeric sequences were removed using usearch61 (Edgar 2010) as implemented in

242 QIIME, with both de novo and reference based detection, and a total of 1,240,976 reads (mean
243 length 203 bp \pm 38 SD) included in the analyses. Short reads representing coral mitochondrial
244 16S or *Symbiodinium* chloroplast 16S reads were removed. Clustering was based on 97%
245 similarity cut-off value and no singletons. Sequences were aligned and classified against the
246 SILVA database (release 123) (Pruesse *et al.* 2007). Chloroplast-assigned reads were removed, as
247 well as rare OTUs (i.e., comprising <0.005% of all sequences), often associated with spurious
248 reads (Bokulich *et al.* 2013). All alpha- and beta-diversity analyses of microbial communities
249 were based on relative abundances at OTU level (97% similarity) and were performed in R
250 (R_Development_Core_Team 2013) using packages ‘vegan’, ‘labdsv’, ‘MASS’, ‘cluster’,
251 ‘indicpecies’, ‘permute’, ‘dplyr’ and ‘tidyr’. Firstly, 800 sequences were randomly picked from
252 each sample in order to correct for different sequencing depths; sequencing produced many host
253 16S mitochondrial reads, as well as 16S chloroplast reads that were removed before analyses,
254 resulting in a low number of prokaryote reads for some samples. Rarefaction curves indicated
255 that even with only 800 sequences, saturation was approached in almost all samples (data not
256 shown). OTU richness and Shannon-Weaver diversity for the different sampling groups were
257 compared by Analysis of Variance (ANOVA). To determine whether habitat and location drive
258 the variation in community assemblage, these factors were included in a Canonical
259 Correspondence Analysis (CCA), which applied an automatic backward and forward model
260 selection tool for constrained ordination methods (‘ordistep’, based on AIC index, with a
261 maximum of 200 permutations) before final significance was determined by an ANOVA-like
262 permutation test based on 9999 permutations. To visualize variation in community composition
263 among habitats and locations, a non-metric multidimensional scaling ordination (nMDS, after
264 9999 permutations) was run based on Bray-Curtis dissimilarities of relative abundance data.
265 Differences in community structure were tested by analysis of similarity (ANOSIM; 9999
266 permutations) and by permutational multivariate analysis of variance (PERMANOVA, 9999
267 permutations) after confirming for homogeneity of multivariate dispersion using a measure of
268 betadispersion (PERMDISP). The OTUs significantly associated with the distinct levels of
269 significant community drivers were detected by Indicator Values (IndVal) analysis, which
270 resolves the taxonomic groups with the highest probability of specificity and fidelity to a
271 particular sampling group.

272

273 *Symbiodinium* community composition

274 PCR and sequencing proceeded as described for the prokaryotic community, with the exception
275 that primers targeting the *Symbiodinium* nuclear DNA ribosomal partial 5.8S, the entire ITS2 and
276 the partial 28S gene (itsD 5'-GTGAATTGCAGAACTCCGTG-3' and its2rev2 5'-
277 CCTCCGCTTACTTATATGCTT-3'; Pochon *et al.* 2001) were used in place of the prokaryote
278 16S V4 primers. Retrieved raw reads plus respective metadata have been deposited in the NCBI
279 Sequence Read Archive under BioProject accession number PRJNA380172. Read assembly,
280 denoising and chimera removal were performed in QIIME (version 1.9.0) (Caporaso *et al.* 2010)
281 as described above. Only assembled reads for which both forward and reverse primers were
282 found were kept in the analysis, comprising a total of 3,453,257 reads with a mean length of 300
283 bp (± 11 SD). To assign taxonomic identity to sequences, we employed a three-step clustering
284 approach in QIIME. Firstly, we performed a rough clustering at 80% similarity cut-off against
285 representative sequences of each recognized *Symbiodinium* clade (A, B, C, D, E, F, G, H and I)
286 retrieved from a database compiled by Arif *et al.* (2014). This database includes 433 distinct ITS2
287 types taken from the GeoSymbio (Franklin *et al.* 2012), Scott Santos'
288 (www.auburn.edu/~santosr/sequencedatasets.htm) and Todd LaJeunesse's SD2-GED
289 (https://131.204.120.103/srsantos/symbiodinium/sd2_ged/database/views.php) databases. This
290 allowed removing non-*Symbiodinium* reads (such as sequencing errors or reads belonging to
291 other dinoflagellate taxa) and assigning 3,449,467 *Symbiodinium* ITS2 sequences to their
292 respective clades. Secondly, clade-assigned sequences were clustered at 97% similarity cut-off
293 (shown to cluster ITS2 intragenomic variants into a single OTU; Arif *et al.* 2014) against all ITS2
294 types available in the Arif database for that particular clade using a closed-reference approach.
295 Finally, the failed assignments of the second step (i.e., not assigned to the reference database)
296 were clustered at 97% similarity with an open-reference OTU picking method using usearch61
297 (Edgar 2010). Resulting OTU tables were then merged into a single summary file representing all
298 ITS2 types found (across all *Symbiodinium* clades), from which singletons and rare OTUs (i.e.,
299 comprising <0.005% of all sequences) were removed. All further statistical analyses were
300 performed on *Symbiodinium* OTUs at 97% similarity following the approach described above for

301 the prokaryotic community after rarefying abundance distributions to 20,000 *Symbiodinium* reads
302 per sample (i.e., the lowest number of reads among the samples). Effects of host genetic structure
303 on *Symbiodinium* communities was tested by PERMANOVA, ANOSIM, and CCA as explained
304 above for the prokaryote 16S rRNA gene data.

305

306 **Results**

307 *Environmental variables*

308 *In situ* measurements at the sites of collection over an 18 months period showed that while the
309 mean seawater temperatures were similar between the reef flat and slope sites, temperature
310 ranges were much larger on the flat compared to the slope (annual range flat: 15.8 - 33.2°C;
311 slope: 18.7° - 28.0°C; Fig. 1A and Supplementary Fig. S1). Photosynthetically active radiation
312 (PAR) varied between the flat and the slope sites and was approximately three times higher and
313 four times more variable on the flat compared to the slope (daily PAR between 9 am and 3 pm at
314 flat: 150 - 1,400 $\mu\text{mol photons.m}^{-2}.\text{s}^{-1}$, and slope \sim 100 - 400 $\mu\text{mol photons.m}^{-2}.\text{s}^{-1}$) (Fig. 1B).
315 Mean of PAR over the period March-August 2013 was 781, 743 and 84 $\mu\text{mol photons.m}^{-2}.\text{s}^{-1}$ at
316 Gardens reef flat, Canyons reef flat and Gardens reef slope, respectively.

317

318 *Coral ID*

319 All samples, except GS203 and GS221, were successfully PCR amplified and RE digested using
320 the protocol of Torda *et al.* (2013c). All of these were confirmed to be *P. damicornis*, although
321 GS209 showed a restriction digestion pattern suggestive of incomplete digestion, and could
322 therefore not be confidently identified.

323

324 *Coral host genetic partitioning*

325 Sequencing of the RAD-seq libraries resulted in an average of 3.6 million reads per sample. A
326 total of 23,350 SNPs in 5,063 RAD-seq loci were identified across 94 *P. damicornis* individuals
327 (two samples failed). We only retained the 2,091 loci that mapped to publicly available
328 *Pocillopora* transcriptomes, and that did not have matches against available *Symbiodinium*
329 genomes and transcriptomes. Additional filtering, as described in the Material and Methods
330 section, reduced this data set to 2,268 SNPs across 890 *Pocillopora* transcriptome loci. We
331 identified 10 sets of highly similar individuals (>94% allelic similarity), likely representing
332 clones. Eight pairs occurred at the same site, and one individual of each clone-pair was removed
333 from the data set (GF105, GF108, GF112, GF120, GS204, GS207, GS209, CS409) as well as
334 three individuals from a clone-set (CS412, CS420, CS423). Where clone mates occurred at
335 different collection sites, both individuals were retained within the data set (CF301 and GF115,
336 CS407 and GS213, GS211 and CS420). This resulted in a non-clonal dataset of 83 individuals.

337 Strong genetic partitioning occurred between habitats rather than between locations (Figs. 2A and
338 B). Under the assumption of two genetic clusters ($K=2$), the Bayesian model-based clustering
339 analysis implemented in STRUCTURE (Fig. 2A) showed a partitioning between reef flat versus
340 slope populations for both the overall and "neutral" datasets, with a small number of individuals
341 with mixed probability of assignment to both genetic clusters, particularly in the reef flat sites.
342 Additional structuring within habitats became apparent for $K=3$ (identified as "optimal K " using
343 both estimator methods for the "neutral" dataset), however for replicate runs this structure
344 occurred alternatively in the flat and slope habitat. At $K=4$ (Fig. 2A), the substructure was
345 identified in both habitats. Although STRUCTURE results for the overall dataset and "neutral"
346 SNPs (2,084 SNPs) were similar, the "neutral" run only identified two admixed individuals: one
347 on the reef flat and one on the reef slope of Coral Gardens (Supplementary Fig. S2). We did not
348 observe any evidence for cross-habitat migrants, and the general lack of admixture (except for the
349 two individuals mentioned above) is indicative of limited gene flow between habitats. Both the
350 genetic partitioning between habitats and further differentiation within habitats were confirmed
351 by a Principal Component Analysis (PCA) (Fig. 2B). Global pairwise F_{ST} distances (i.e., genetic
352 differentiation based on all SNPs) between habitats were roughly double that between locations
353 (Coral Canyons and Coral Gardens) (Supplementary Table S1).

354 Our conservative outlier analysis, which only included SNPs that were outliers in both replicate
355 slope - flat comparisons and were not identified as outliers between Coral Canyons and Coral
356 Gardens, revealed 16 outlier SNPs in 13 loci (Fig. 3). The loci were mapped back to the longer
357 contigs in the publicly available *Pocillopora* transcriptomes and those contigs were annotated
358 using BLASTx in ZoophyteBase (Dunlap *et al.* 2013). Ten of the outlier loci had a robust
359 annotation based on e-values ($<10^{-8}$) and bit scores (Table 1), four of which match genes
360 involved in the innate immune response, four are environmental stress response genes, one is
361 involved in lipid metabolism and one plays a role in DNA processing. Battenin is a protein
362 encoded by the *CLN3* gene that is involved in the lysosome system. Lysosomes are membrane-
363 bound organelles that contain hydrolytic enzymes, including lysozyme, which is regarded as one
364 of the key innate immune effector molecules against bacteria by causing bacterial lysis. NF- κ B is
365 a 'master' immunity regulator (Davy *et al.* 2012). The NF- κ B inhibitor-like protein 2 is a
366 transcription factor that regulates the innate immune response; it acts as negative regulator of the
367 Toll-like receptor and interferon-regulatory factor signaling pathways. As such, it contributes to
368 the negative regulation of transcriptional activation of *NF- κ B* target genes in response to
369 endogenous proinflammatory stimuli. Cannabinoid receptor 2 is involved in the release of nitric
370 oxide (NO). NO is a direct signaling molecule between *Symbiodinium* and coral host that may
371 initiate a bleaching cascade (Weis 2008). Histamine receptor H2 is an integral membrane protein
372 and mediates the actions of histamine; histamine is involved in local immune responses to foreign
373 pathogens. Ubiquitin-protein ligases (*E3*, *UBR3*, *UBR5*, *EDD1*) and ubiquitin thioesterase
374 (*CYLD*, *USLP2*) have functions in the ubiquitin system which is responsible for the selective
375 degradation of proteins. ApaH, bis (5'-nucleosyl)-tetrphosphatase hydrolyzes diadenosine 5',5''-
376 P1,P4-tetrphosphate, a regulatory metabolite of stress conditions, to yield ADP.

377

378 *Coral tissue-associated prokaryotes*

379 PCR using primers targeting the 16S rRNA V4 region yielded two amplicons (~330 and ~410 bp
380 in size) in three of the four populations, and only the shorter product in the reef slope population
381 of Coral Gardens. Sequence analysis indicated this shorter product was of coral mtDNA origin,

382 therefore, no prokaryote 16S rRNA gene data were obtained for this population. Further, the
383 prokaryote communities in the water column above the sampled coral colonies were not
384 examined and it can therefore not be ruled out that coral-associated communities were shaped by
385 differences in the communities present in the water column in the two habitats. We consider this
386 explanation unlikely given that connectivity among water-borne microbiomes separated by <100
387 m distance is likely high (Frade *et al.* 2016; Martiny *et al.* 2006). Nevertheless, the differences in
388 prokaryote community composition between reef flat and slope habitats we describe below need
389 to be taken with some caution. Samples from the remaining three sites yielded between 120 and
390 20,187 prokaryotic reads. Samples with <800 reads (five samples) were removed, leaving 35 flat
391 and 20 slope samples with a mean number of reads of $7,348 \pm 4630$ (SD) and a median of 6,691
392 for further analyses. Taxonomic assignment (see Supplementary Fig. S3 for relative abundances)
393 revealed that operational taxonomic unit (OTU) diversity was low (~27 OTUs per sample, range
394 = 13-55) compared to other studies that show 100s-1000s of OTUs per colony (Blackall *et al.*
395 2015; Bourne *et al.* 2016; Hernandez-Agreda *et al.* 2016a). Rarefaction at 800 reads per sample
396 will have contributed to the low diversity observed. Further, excessive production of mucus
397 during collection in the field would have diluted the mucus-associated microbial community, and
398 EtOH preservation of the samples removed some of the coral mucus where a taxonomically
399 diverse community of prokaryotes resides. Hence, the OTUs identified ($n = 200$) here are likely
400 associated with the coral tissue.

401 The reef flat was significantly less rich in OTUs than the reef slope (observed OTU richness of
402 23 ± 6 and 34 ± 9 , respectively; $F_{(53,1)} = 32.88$, $p < 0.001$), and also less diverse (Shannon-Weaver
403 diversity of 1.18 ± 0.41 and 1.64 ± 0.26 for reef flat and reef slope, respectively; $F_{(53,1)} = 20.95$,
404 $p < 0.001$), consistent with previous findings for the related coral species *Seriatopora hystrix*
405 (Pantos *et al.* 2015). When non-rarefied data were used, the number of OTUs resolved increased
406 to 38 ± 9 for the flat, and 78 ± 11 for the slope, an approximately two-fold increase in richness
407 compared to the rarefied data. Shannon diversity was stable at 1.19 ± 0.40 and 1.67 ± 0.26 ,
408 respectively for flat and slope, showing that by rarefying we only excluded very rare members of
409 the microbiome (likely with uneven distributions). No differences in alpha-diversity were found
410 between locations.

411 As with the host genome data, ordination analysis of the prokaryotic community assemblages
412 (beta-diversity) showed a clear division between habitats (i.e., reef flat and slope) at the Coral
413 Canyons where this could be assessed (Fig. 4; Supplementary Figs. S4A - D). Differences
414 between reef habitats were statistically supported by two distinct approaches (PERMANOVA
415 pseudo $F_{(53,1)} = 34.402$; $p < 0.001$; and ANOSIM statistic $R = 0.6042$, $p < 0.001$). Constrained beta-
416 diversity comparisons via Canonical Component Analysis (CCA) showed that both habitat ($F_{(53,1)} = 6.5572$, $p < 0.001$) and location ($F_{(53,1)} = 1.6006$, $p = 0.0323$) were significant drivers of
417 prokaryotic community composition. Non-overlapping distributions of reef flat samples
418 belonging to the two distinct locations in the non-metric multidimensional scaling (nMDS)
419 ordination (Fig. 4) supports that in addition to partitioning between habitats, differences in
420 prokaryote communities exist between the locations: Coral Canyons and Coral Gardens. The
421 non-homogeneous multivariate dispersion verified among reef locations (PERMDISP $F_{(53,1)} =$
422 13.466 ; $p < 0.001$) precluded any further statistical comparison based on dissimilarity
423 distributions. Host population genetic structure (based on $K = 4$ in the STRUCTURE analyses)
424 within habitats did not explain any additional variation (Supplementary Figs. S4C, D).

426 The difference in prokaryotic community composition between the reef flat and slope populations
427 of *P. damicornis* was driven by a dominance of *Endozoicomonas* sp. (Gammaproteobacteria,
428 Oceanospirales, Hahellaceae) on the flat, while the slope population showed co-dominance of
429 *Endozoicomonas* sp. and *Brevibacterium* sp. (Actinobacteria, Micrococcales, Brevibacteriaceae).
430 Other taxa were present mostly below 1% relative abundance (Archaea $< 0.1\%$ relative
431 abundance) (Supplementary Fig. S3). Twenty-three OTUs were significantly associated with a
432 particular habitat (Fig. 5), ten of which belonged to *Brevibacterium* sp. and which were
433 significant indicators (Glasl *et al.* 2017b) of the reef slope community. The most abundant of
434 these OTUs had an affinity to *Brevibacterium marinum* (AM421807; Lee 2008), reaching on
435 average 28.5% of the reads retrieved from the reef slope samples, compared to only 2.2% on the
436 flat. This was also the most dominant OTU within the genus, with an average relative abundance
437 of $> 96\%$ of the *Brevibacterium* sp. reads across habitats and locations (Supplementary Fig. S4A).
438 Two low-abundance *Endozoicomonas*-affiliated OTUs were also identified as indicators for the
439 reef slope, despite the fact that *Endozoicomonas* spp. were overall more abundant on the flat. The

440 most dominant *Endozoicomonas* OTU (KC669021; Bayer *et al.* 2013) was ubiquitous and
441 represents >73% of the reads within the genus across all samples (Supplementary Fig. S4B).
442 Other indicator taxa for the reef slope (Fig. 5) included a *Nevskia* sp. (Gammaproteobacteria,
443 Xanthomonadales, Nevskiaceae), *Halomonas salina* (Gammaproteobacteria, Oceanospirillales,
444 Halomonadaceae), a *Bradythizobium* sp. (Alphaproteobacteria, Rhizobiales, Bradyrhizobiaceae),
445 a *Dietzia* sp. (Actinobacteria, Corynebacteriales, Corynebacteriaceae) and a *Corynebacterium* sp.
446 (Actinobacteria, Corynebacteriales, Corynebacteriaceae). The only OTU pinpointed as a
447 significant indicator for the reef flat was a *Tenacibaculum* sp. (Bacteroidetes, Flavobacteriales,
448 Flavobacteriaceae).

449

450 *Intracellular photosymbionts in the genus Symbiodinium*

451 Eighty-nine samples across the four sampling sites yielded between 21,179 and 56,014
452 *Symbiodinium* ITS2 reads, with a mean number of reads of $37,806 \pm 7567$ (SD) and a median of
453 37,417. These corresponded to 55 distinct OTUs (at the 97% sequence similarity cut-off),
454 including 45 clade C types and one clade B type assigned from the Arif *et al.* database (Arif *et al.*
455 2014), and nine new OTUs (Fig. 6). Compared to the reef slope, the reef flat showed higher
456 *Symbiodinium* richness (44 ± 2 against 35 ± 5 OTUs; $F_{(87,1)} = 100.7$, $p < 0.001$) and Shannon-
457 Weaver diversity (1.35 ± 0.09 against 0.75 ± 0.22 ; $F_{(87,1)} = 291.9$, $p < 0.001$). ITS2 sequence
458 types C33 and C42 were dominant on the reef slope and reef flat, respectively, together with
459 other sequence types with matches to C1, C1.8 and C23 (Fig. 6). It should be noted that our
460 approach could not distinguish C42 from C42a (as they have identical sequences in the Arif *et al.*
461 database for the region sequenced in this study) or C33 from C33a (as divergence is too low to be
462 captured by the similarity cut-off value applied). However, when clustering was performed at a
463 99% sequence similarity cut-off, the majority of C33-associated sequences were reassigned to
464 C33a, with only a minor fraction (5-15%) of those sequences assigned to C33. Because
465 sequences types C33a, C33 and C42 have previously been identified in *P. damicornis* colonies
466 from the slope at a Heron Island reef nearby our collection sites (Sampayo *et al.* 2007), we
467 assume the same *Symbiodinium* types are present in the colonies from Coral Canyons and Coral

468 Gardens. Differences in *Symbiodinium* community structure between the reef flat and slope (Fig.
469 6) were confirmed by PERMANOVA (pseudo $F_{(87,1)} = 807.16$; $p < 0.001$) and ANOSIM statistic
470 ($R = 0.9817$, $p < 0.001$) after verifying for homogeneous multivariate dispersions between the two
471 habitats. Constrained beta-diversity comparisons (via CCA) showed that only habitat ($F_{(85,1)} = 215.7771$,
472 $p < 0.001$) was a significant driver of *Symbiodinium* community assemblage.
473 Location was not a significant structuring factor in *Symbiodinium* alpha-diversity or community
474 structure parameters, with most samples within each habitat having near identical relative
475 abundances of *Symbiodinium* sequences. Host population genetic structure within sites also did
476 not explain variation in the *Symbiodinium* communities, and there was no correlation between
477 *Symbiodinium* and prokaryote community composition (Fig. S5).

480 Discussion

481 This study demonstrates genome-wide differentiation between adjacent populations of the reef-
482 building coral, *Pocillopora damicornis*, which inhabit the reef flat and slope at Heron Island in
483 the southern GBR. Of the 2,268 SNPs included in the analyses, ~0.7% were conservatively
484 identified as outliers between the flat and slope habitats, indicating the importance of
485 environmental selection on the coral host genome. Community composition of intracellular
486 photosymbionts (*Symbiodinium* spp.) and coral tissue-associated prokaryotes also differed
487 significantly between the reef flat and slope populations of *P. damicornis*. Few previous studies
488 have examined divergence between habitats in the coral host genome (rather than host genotype
489 based on one or a few neutral markers) simultaneously with the associated *Symbiodinium*
490 communities (Barshis *et al.* 2013; Bongaerts *et al.* 2017); only two have examined all three
491 compartments (coral host, *Symbiodinium* and prokaryotes) at the same time, but no functional
492 analysis of loci under selection was conducted (Gonzalez-Zapata *et al.* 2018; Pantos *et al.* 2015).
493 Our novel findings indicate that local adaptation to the distinct reef habitats, flat and slope, has
494 occurred through selection on at least three members of the coral holobiont: the coral host, its
495 *Symbiodinium* and the tissue-associated prokaryotes.

497 *Signatures of local adaptation in the coral genome*

498 The majority of the outlier genes identified for *P. damicornis* from the reef flat and slope at
499 Heron Island are involved in immune and environmental stress responses (Table 1). High
500 temperature and light levels, as well as a number of other stressors, cause an increased generation
501 of Reactive Oxygen Species (ROS) and Reactive Nitrogen Species (RNS) by the coral
502 photosymbionts, *Symbiodinium* spp., and to some extent by the coral host mitochondria (Lesser
503 2006; Lesser 2011; Weis *et al.* 2008). High ROS levels can stimulate innate immune-like
504 pathways in the coral host (Hawkins & Davy 2012; Perez & Weis 2006) that are part of the
505 conserved minimal cellular stress response (Kültz 2005). Certain components of the stress
506 response are evolutionarily highly conserved and recruited by a number of different stressors
507 (Kültz 2003). For example, in marine oysters it has been demonstrated that both biotic
508 (pathogens) and abiotic (e.g., acidified water) stressors affect mitochondrial function, the
509 cytoskeleton, energy production and ROS generation. These stressors cause the recruitment of the
510 same cellular pathways (Anderson *et al.* 2015) involving genes encoding antioxidant enzymes
511 (e.g., superoxide dismutases - *SODs*), chaperones (e.g., heat shock proteins - *HSPs*), removal of
512 macromolecular debris generated by stress (ubiquitin/proteasome pathway), the *NF- κ B* pathway
513 and genes involved in apoptosis and cytoskeleton remodelling (Kültz 2003). Transcriptome
514 analysis in the damselfish, *Acanthochromis polyacanthus*, showed immune- and stress-response
515 as well as metabolic genes underpin transgenerational acclimation to elevated temperature
516 (Veilleux *et al.* 2015). The coral heat-stress response is also known to involve chaperones, such
517 as heat shock proteins, and the ubiquitin system (Csaszar *et al.* 2009; Desalvo *et al.* 2008;
518 DeSalvo *et al.* 2010; Maor-Landaw *et al.* 2014; Voolstra *et al.* 2009). Further, a growing body of
519 evidence supports local adaptation to different environmental factors, including temperature,
520 involves selection on genes in the ubiquitin system (Barshis *et al.* 2010; Bay & Palumbi 2014; Jin
521 *et al.* 2016; Lundgren *et al.* 2013; Thomas *et al.* 2017). Our results on *P. damicornis* are
522 consistent with previous findings and highlight the prevalent role of the ubiquitin system in the
523 coral acclimatization and adaptive responses to environmental differences.

524

525 *Selection on coral-associated microbial communities*

526 While we provide compelling evidence for local environmental selection on the coral host
527 genome, the differences in the community composition of *Symbiodinium* and prokaryotes hosted
528 by slope and flat coral populations suggest the coral microbiome also plays important roles in
529 local adaptation and the ability of corals to adjust to environmental differences and change.
530 *Symbiodinium* communities associated with *P. damicornis* colonies located between 3 and 19 m
531 depth at Heron Island have previously been assessed using Denaturing Gradient Gel
532 Electrophoresis analysis of the ITS2 region (Sampayo *et al.* 2007) and sequence analysis of a the
533 ITS2 and other regions (Sampayo *et al.* 2009). However, this earlier study only included samples
534 from the reef slope and no reef flat colonies were examined. The spatially closest sampling
535 location in these previous studies to our reef locations is Harry's Bommie, ~600 m away from
536 Coral Canyons. The *P. damicornis* colonies at Harry's Bommie were dominated by a
537 combination of ITS2 sequence types C33 and C33a, consistent with our results for the slope
538 populations. Sequence types C42 and C1 were also present in our and the previous studies, and,
539 as expected for the higher resolution amplicon next generation sequencing method, we observed
540 additional sequence types previously reported from cloning methods (Sampayo *et al.* 2009; Fig.
541 6). These co-occurring sequence types likely represent single but distinct *Symbiodinium* types in
542 *P. damicornis* colonies on the reef flat versus the slope. Conspecific coral colonies that are
543 dominated by distinct *Symbiodinium* types can differ considerably in physiological performance
544 and environmental tolerance (Blackall *et al.* 2015; van Oppen *et al.* 2009), and we hypothesize
545 that adaptation to the reef flat and slope environments by *P. damicornis* at Heron Island is partly
546 contingent on the distinct *Symbiodinium* communities harboured.

547 Community composition of prokaryotes associated with the *P. damicornis* tissues differed
548 between the slope and flat at the location where this could be assessed, with one and 22 indicator
549 OTUs being identified for the flat and slope respectively. Additional genetic structure was
550 present between locations but this divergence was smaller than that between habitats. Other
551 prokaryotic taxa associated with the coral surface mucus layer may also differ between the two

552 habitats, but these organisms were not examined here. The corals inhabiting the reef flat were
553 dominated by *Endozoicomonas* spp. (Gammaprotobacteria, Oceanospirales, Hahellaceae), while
554 the slope population showed co-dominance of *Endozoicomonas* sp. and *Brevibacterium* sp.
555 (Actinobacteria, Micrococcales, Brevibacteriaceae). *Endozoicomonas* spp. are commonly
556 associated with reef-building corals, and can aggregate inside the tissue of corals in the family
557 *Pocilloporidae* (Bayer *et al.* 2013; Blackall *et al.* 2015; Morrow *et al.* 2015; Morrow *et al.* 2012;
558 Neave *et al.* 2016a; Neave *et al.* 2016b; Pantos *et al.* 2015). The most dominant *Endozoicomonas*
559 OTU matches an OTU associated with another pocilloporid, *Stylophora pistillata*, from the Red
560 Sea (Bayer *et al.* 2013). Another OTU is a perfect match to an *Endozoicomonas* discovered in *P.*
561 *damicornis* colonies from the Red Sea (KC668747; Bayer *et al.* 2013). These findings are
562 concordant with recent observations of strong specificity of certain *Endozoicomonas* OTUs in
563 *Pocillopora verrucosa*, although the authors argued that this may be related to their broadcast
564 spawning behaviour (Neave *et al.* 2016b); *P. damicornis* is primarily an asexual brooder, but has
565 been observed to also release gametes (Schmidt-Roach *et al.* 2012). *Brevibacterium* spp. are less
566 well known from corals. Nevertheless, *Brevibacterium aureum* was present at a mean relative
567 abundance of ~31% in all *Porites lobata* colonies sampled from the Persian/Arabian Gulf
568 (Hadaidi *et al.* 2017). *Brevibacterium* was also found to be the dominant taxon among culturable
569 bacteria isolated from the mucus and tissue of three coral species from the Arabian Gulf
570 (Mahmoud & Kalendar 2016).

571 Our findings of strong habitat differentiation in the prokaryotic communities of a depth generalist
572 (*P. damicornis* occurs from the reef flat down to ~55 m of depth; Bridge *et al.* 2012) align with
573 those of Glasl *et al.* (2017a), who concluded that corals with a relatively broad depth distribution
574 show structuring in tissue-associated prokaryotic communities across their depth range.
575 Environmental factors can drive coral-associated prokaryotic community composition
576 (Hernandez-Agreda *et al.* 2016b; Littman *et al.* 2009; Pantos *et al.* 2015), and may explain the
577 patterns observed in *P. damicornis* at Heron Island. Alternatively, the differences in prokaryote
578 community composition between the flat and slope populations may be caused by both the coral
579 host and the *Symbiodinium* photosymbiont having undergone divergence and niche
580 diversification (Bongaerts *et al.* 2015; Bongaerts *et al.* 2013; Bongaerts *et al.* 2010). This could

581 in turn has led to the existence of particular holobiont microniches to which at least a small
582 fraction of the microbial community is adapted (Glasl *et al.* 2017a). The challenge now is to
583 determine how the different prokaryote communities affect the functioning of the coral holobiont
584 across these different habitats, and the extent to which adaptation of the coral host and its
585 microbiome are linked.

586

587 *Implications for human intervention strategies that aim to enhance coral climate resilience*

588 The GBR off the east coast of Australia spans 14 degrees of latitude and a mean sea surface
589 temperature (SST) gradient of $\sim 3^{\circ}\text{C}$, hence, corals in the northern GBR have evolved to cope
590 with higher temperatures compared to their southern counterparts (Berkelmans 2002; Berkelmans
591 & Willis 1999; Howells *et al.* 2013). Translocation of corals from warm to cooler locations (i.e.,
592 assisted gene flow) has been suggested as a strategy to increase thermal tolerance in the cooler
593 locations (Hoegh-Guldberg *et al.* 2008; Weeks *et al.* 2011), especially if the translocated colonies
594 breed with the corals native to the cooler but warming environment (Dixon *et al.* 2015; van
595 Oppen *et al.* 2014). When corals from the northern (warmer) and central (cooler) GBR were
596 crossed in the laboratory, their regional hybrid larval offspring showed higher heat tolerance
597 under laboratory conditions than purebreds from the cooler location, especially if the mother was
598 sourced from the warmer location (Dixon *et al.* 2015). These results indicate that north to south
599 translocation along the GBR may be a valuable strategy to assist southern GBR coral populations
600 to adapt to a warming ocean. A common concern regarding translocations between different
601 locales is the risk of introducing pathogens or parasites from the donor into the receiving
602 population, negatively affecting the health of the latter (Laikre *et al.* 2010). Reduced resistance of
603 migrants to local pathogens has also been reported (Grady *et al.* 2015; Van der Putten 2012).
604 Further, translocated adult colonies may be maladapted to other environmental conditions at the
605 new locations and may not survive to reproduce (Howells *et al.* 2013). A safer alternative may
606 therefore be to harness the adaptive variation contained in distinct temperature habitats on the
607 same reef. Our results demonstrate that coral populations from the reef flat and slope can be
608 genetically divergent and adapted to their local environments. The limited gene flow between

609 slope and flat populations is likely due to maladaptation to environmental conditions rather than
610 physical restrictions to larval dispersal given the small spatial distances between habitats (<100
611 m) (Marshall *et al.* 2010). Assuming these populations are still cross-fertile, artificial *ex-situ*
612 habitat-crosses may provide an avenue to breed increased thermal tolerance into conspecific
613 slope populations, the reef habitat where coral cover tends to be highest. Inoculations with
614 microbiomes (probiotics) (Damjanovic *et al.* 2017; Peixoto *et al.* 2017) isolated from the reef flat
615 corals may be another avenue that can be explored to increase thermal resilience in the slope
616 populations.

617

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938 Data Accessibility

939 RADseq data are available as fastq files on NCBI Sequence Read Archive (accession number
940 PRJNA450871), and VCF files on Dryad (doi:10.5061/dryad.j09c05c). Prokaryote community
941 raw reads plus respective metadata have been deposited in the NCBI Sequence Read Archive
942 under BioProject accession number PRJNA380169. Raw reads for the *Symbiodinium* community
943 data plus respective metadata have been deposited in the NCBI Sequence Read Archive under
944 BioProject accession number PRJNA380172.

945

946 Author Contributions

947 M.v.O. and L.B. conceived and designed the study, conducted fieldwork and collected samples;
948 L.P. conducted all laboratory work; P.B., P.F., S.B., H.N. and M.v.O. performed statistical
949 analyses and interpreted the results; M.v.O., P.B., P.F. and L.B. wrote the manuscript and all
950 authors edited various versions of the manuscript.

951 Figure Captions:

952

953 Figure 1: *In situ* measurement of environmental variables. (A) Temperature in degrees Celcius
954 over two years. The habitat temperatures were indistinguishable between the two locations and
955 only Coral Gardens is displayed (Slope in blue and Flat in red). (B) PAR in $\mu\text{mol photons}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$
956 measured over eight months in 2013. The light levels in habitats were indistinguishable between
957 locations and only Coral Gardens is displayed (Slope in blue and Flat in red).

958 Figure 2: Coral host genetic structure based on all SNP loci. (A) STRUCTURE results at K=2
959 and K=4 with individuals on the x-axis (sorted by population) and assignment probability on the
960 y-axis. (B) PCA. Colors in (B) indicate the cluster assignment in (A). Grey color indicates
961 individuals that had a STRUCTURE assignment of <0.8 . Circles indicate reef flat sites, triangles
962 indicate reef slope sites as shown in (A).

963 Figure 3: Coral host genetic differentiation and outlier loci. (A) Genetic differentiation versus
964 expected heterozygosity, with a histogram on the y-axis showing the F_{ST} frequency distribution
965 for all SNPs. Colors indicate the different outlier categories (as described in the text). (B)
966 Genotype frequencies of the 16 habitat-outlier SNPs, with the hues indicating the frequency of
967 homozygote (ref = reference allele, alt = alternate allele) and heterozygote (ref/alt) genotypes.

968 Figure 4: Ordination analysis of coral tissue-associated prokaryote beta-diversity based on 16S
969 rRNA gene sequence data (nMDS plot) in the two reef flat and the one reef slope populations.
970 CF= Coral Canyons Reef Flat, CS=Coral Canyons Reef Slope, GF= Coral Gardens Reef Flat.

971 Figure 5: Bubble plots of significant indicator prokaryotic OTUs in the two habitats, reef flat and
 972 reef slope. Bubble sizes are representative of average relative abundance across all samples of
 973 each habitat. Codes “d”, “p”, “c”, “o”, “f”, “g” and “sp” represent OTU taxonomic affiliation at
 974 domain, phylum, class, order, family, genus and species level, respectively. Entries between
 975 brackets represent NCBI Accession Numbers for known representative sequences of each OTU
 976 (or “new” in case no match was found above 97% similarity in the SILVA database)

977 Figure 6: Relative abundance of the ten most abundant *Symbiodinium* ITS2 sequence types (plus
 978 ‘others’ shown as one category) at the two reef flat and two slope sites. CF= Coral Canyons Reef
 979 Flat, CS= Coral Canyons Reef Slope, GF= Coral Gardens Reef Flat, GS= Coral Gardens Reef
 980 Slope.

981

982 Table 1: Annotation of outlier loci based on BLASTx in ZoophyteBase. Dark grey background
 983 indicates genes involved in the immune response, light grey background indicates genes involved
 984 in environmental stress responses, white background indicates ‘other’.

Contig name	Annotation of top BLASTx hit	Putative function (KEGG)	E-value	Bit score	Query cover (%)
427	Battenin (BTS, CLN3)	Cellular Processes: Lysosome system	3.5e ⁻⁴⁶	161.4	50.1
2762 ^a	5-hydroxytryptamine receptor 7	Environmental Information Processing: Ligand-receptor interaction	1.0e ⁻⁸⁰	260.4	97.1

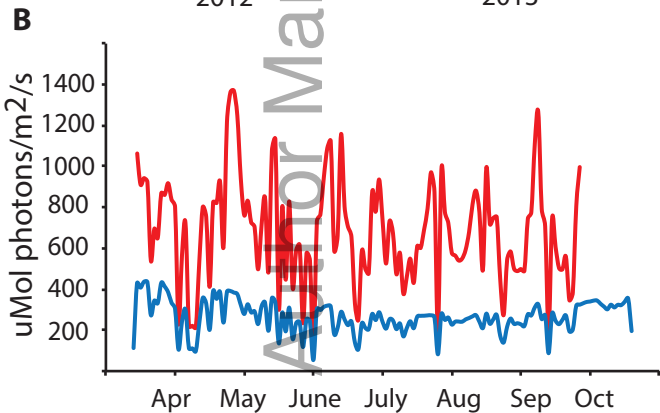
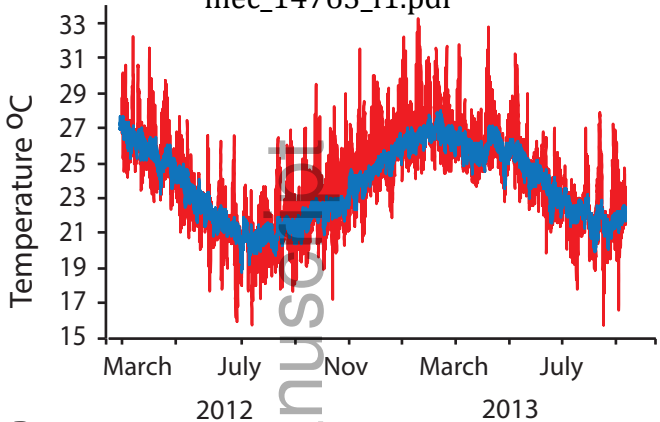
4435	HRH2, histamine receptor H2	Environmental Information Processing: Ligand-receptor interaction, involved in local immune responses to foreign pathogens	$9.3e^{-11}$	58.9	27.9
4491	NF- κ B inhibitor-like protein 2	Genetic information processing: Transcription factor in innate immune response	$2.2e^{-24}$	83.2	11.9
3837	apaH, bis(5'-nucleosyl)-tetraphosphatase	Metabolism: regulatory metabolite of stress conditions	$2.8e^{-9}$	56.2	62.4
804	Ubiquitin-protein ligase (UBR3, E3)	Genetic information processing: Ubiquitin system	$9.5e^{-58}$	200.7	89.1
28	Ubiquitin-protein ligase (EDD1, UBR5, E3)	Genetic information processing: Ubiquitin system	0	1004.2	39.2
2255	Ubiquitin thioesterase (CYLD,	Metabolism:	$8.0e^{-8}$	53.1	6.6

	USLP2)	Ubiquitin system			
1133	ATP-dependent DNA helicase (PIF1)	Genetic information processing: DNA replication proteins	9.9 e ⁻⁵⁸	206.8	13.37
3046	Ectonucleotide pyrophosphatase/phosphodiesterase family member 6 (ENPP6)	Metabolism: Lipid Metabolism	0	562.0	89.8

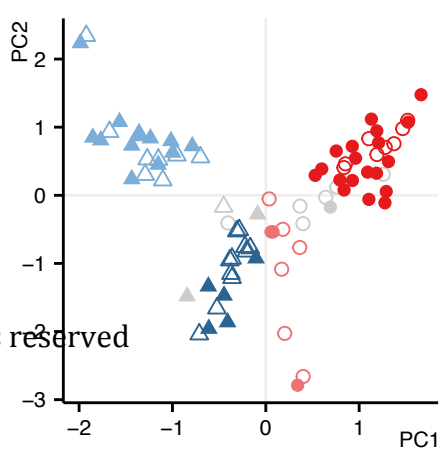
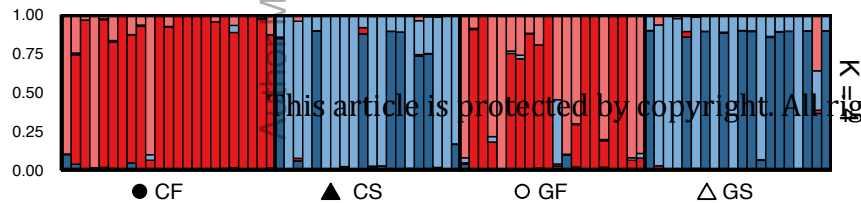
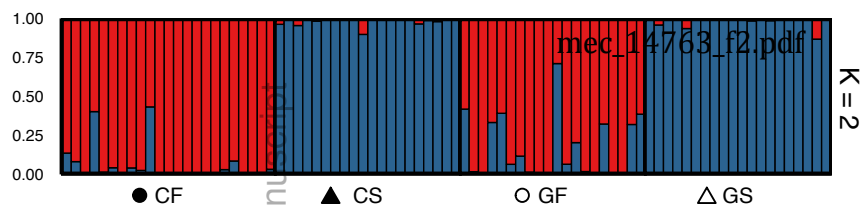
985 ^aSecond best match was to cannabinoid receptor 2 (E-value 1.0e⁻²⁶, bit score 112.5, query cover 76.3

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Date



Genetic differentiation (F_{ST})

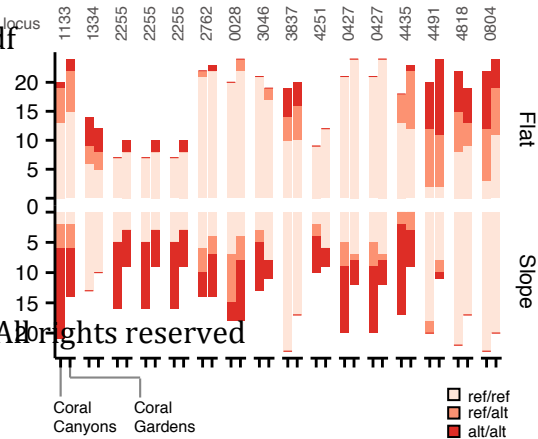
Heterozygosity (H_E)

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- Habitat-outlier
- Fdist & BayeScan
- Fdist
- Non-outlier

Frequency

Genotype frequency



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▲ CS
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